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**TRATAMENTOS SILVICULTURAIS EM CLAREIRAS APÓS EXPLORAÇÃO
FLORESTAL**

**BELÉM
2020**

RAPHAEL LOBATO PRADO NEVES

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FLORESTAL**

Defesa de tese apresentada à Universidade Federal Rural da Amazônia como parte das exigências para obtenção do título de Doutor em Ciências Florestais, área de concentração Manejo de Ecossistemas Florestais.

Orientador: Dr. Gustavo Schwartz

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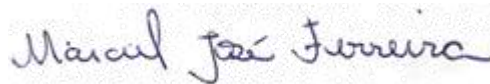
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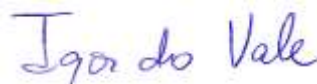
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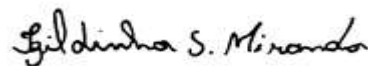
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RESUMO

O objetivo geral do trabalho foi o de avaliar qual o efeito a médio prazo dos tratamentos silviculturais pós-colheita na sobrevivência, crescimento e estrutura de espécies comerciais, tanto de árvores plantadas quanto as de regeneração natural, ambas conduzidas em clareiras de exploração florestal. Essas clareiras foram abertas nos anos de 2004 e 2006, totalizando 72 clareiras sob diferentes tipos de tratamentos com avaliação ao longo de onze anos na área de manejo florestal da empresa Jari Florestal S.A, situada em Monte Dourado, distrito da cidade de Almerim, Amazônia Oriental, Brasil. Foi avaliado a regeneração natural sem aplicação de nenhum tipo de tratamento silvicultural (SRIL); condução da regeneração natural (TNER); condução de espécies plantadas com e sem remoção prévia de resíduos provenientes da colheita de madeira (EP's). No primeiro artigo foi avaliado a sobrevivência, crescimento e estrutura. Os resultados indicaram maior sobrevivência e crescimento de árvores com tratamento silvicultural pós-colheita quando comparado ao tratamento controle, que reflete os procedimentos padrão de exploração de impacto reduzido. No segundo artigo foi avaliado o efeito do tratamento silvicultural pós-colheita para a espécie *Dinizia excelsa*, uma das mais importantes espécies comerciais da Amazônia e que possui baixa densidade natural. Foi comparado a sobrevivência e o crescimento de árvores de *D. excelsa* em diferentes tamanhos de clareiras, tratamentos e classes de exposição de copa. Foi constatado que não há diferença no crescimento dos indivíduos de acordo com o tamanho da clareira, mas de acordo com a classe de exposição de copa. Apesar da alta mortalidade de indivíduos plantados, estes obtiveram melhor desempenho no crescimento, principalmente aqueles que receberam mais luz solar. A *D. excelsa* é uma espécie que responde muito bem, seja para plantio ou condução da regeneração natural em clareiras. No terceiro artigo buscou-se avaliar a espécie *Tachigali glauca*, também uma importante espécie comercial da Amazônia. O tachi é uma espécie que respondeu muito bem ao tratamento silvicultural, principalmente quando há luz do sol disponível em toda sua copa. Obteve ótimos resultados em sobrevivência e incremento sendo esta espécie recomendada para plantio em clareiras. As clareiras de impacto reduzido ou as clareiras naturais que possuam um mínimo de 200 m² são espaços florestais com alto potencial produtivo e/ou conservacionista. Recomenda-se o uso destes espaços aliado aos tratamentos silviculturais de condução tanto de indivíduos plantados quanto da regeneração natural, auxiliando assim em uma mais rápida produção madeireira associada com a conservação de espécies comerciais.

PALAVRAS-CHAVE: condução da regeneração natural; enriquecimento de clareiras; luz solar; tamanho de clareiras; silvicultura tropical.

ABSTRACT

The general objective of the work was to evaluate the medium-term effect of post-harvest silvicultural treatments on the survival, growth and structure of commercial species, both from planted trees and those from natural regeneration, both conducted in canopy logging gaps. These gaps were opened in 2004 and 2006, totalling 72 gaps under different types of treatments with assessment over 11 years in the forest management area of the company Jari Florestal SA, located in Monte Dourado, district of the city of Almerim, Eastern Amazon, Brazil. Was evaluated natural regeneration without any silvicultural treatment, so-called standard procedures of reduced impact-logging (SRIL); tending of natural established regeneration (TNER); tending of enrichment planted species with and without prior removal of residues from wood harvesting (EP's). In the first article, survival and growth were evaluated. The results indicated greater survival and growth of trees with post-harvest silvicultural treatment when compared to the control treatment, which reflects the standard procedures of reduced impact-logging. Legally adopted in Brazil. In the second article, the effect of post-harvest silvicultural treatment for *Dinizia excels* specie, one of the most important commercial species in the Amazon and which has a low natural density, was evaluated. The survival and growth of *D. excelsa* trees in different gaps sizes, treatments and crown exposure classes were compared. It was found that there is no difference in the growth of individuals according to the gap sizes, but according to the crown exposure class. Despite the high mortality of planted individuals, they performed better in growth, especially those who received more sunlight. *D. excelsa* is a species that responds very well, whether for planting or conducting natural regeneration in gaps. The third article sought to evaluate the species *Tachigali glauca*, also an important commercial species in the Amazon. *T. glauca* is a specie that responded very well to silvicultural treatment, especially when sunlight is available throughout its crown. It obtained excellent results in survival and growth and this specie is recommended for enrichment planting in gaps. Gaps that have a minimum of 200 m² are forest spaces with high productivity and / or conservation potential. It is recommended to use these spaces together with the silvicultural treatments for tending both planted individuals and natural regeneration, thus assisting a faster timber production associated with the conservation of commercial species.

KEYWORDS: natural regeneration tending; enrichment planting; sunlight; size of logging gaps; tropical forestry.

CONTEXTUALIZAÇÃO

Manejo Florestal Sustentável é a administração da floresta para obtenção de benefícios econômicos, sociais e ambientais, respeitando-se os mecanismos de sustentação do ecossistema objeto do manejo e considerando-se, cumulativa ou alternativamente, a utilização de múltiplas espécies madeireiras, de múltiplos produtos e subprodutos não-madeireiros, bem como a utilização de outros bens e serviços florestais. No entanto, nem todos os mecanismos de sustentabilidade das florestas manejadas são adequadamente abordados pelos gestores e pela atual legislação florestal brasileira (BRANDÃO et al., 2018). Portanto, o desafio para a silvicultura tropical é garantir níveis adequados de estoque em crescimento e facilitar a regeneração de espécies de valor comercial, além de serviços ambientais.

O Brasil assumiu compromissos nacionais e internacionais de restaurar e reflorestar, pelo menos, 12 milhões de hectares de florestas até 2030 por meio de sua Contribuição Nacionalmente Determinada (CND), a qual foi apresentada no âmbito da Convenção-Quadro das Nações Unidas sobre Mudança do Clima (Brasil, 2015). Essa é parte da estratégia nacional adotada para reduzir as emissões de gases de efeito estufa (GEE) e promover atividades de menor impacto ambiental (VALLE et al., 2020). Uma destas metas é a criação de mecanismos eficientes que garantam o cumprimento da Lei de Proteção de Vegetação Nativa (BRASIL, 2012), com a recuperação de passivos de Áreas de Preservação Permanente (APPs) e RL (VALLE et al., 2020).

O Governo do estado do Pará instituiu através do decreto de número 941 de 03/08/2020 o Plano Estadual Amazônia Agora (PEAA). O plano considera o incremento de cobertura vegetal secundária para contabilidade das remoções estimadas de GEE, na qual objetiva: §1º ter regeneração da vegetação correspondendo a 5,65 milhões de hectares até o ano de 2030; e/ou §2º ter regeneração de cobertura vegetal correspondendo a 7,41 milhões de hectares para o ano de 2035, caso a implementação do PEAA disponha de recursos externos até 2030.

Parte da recuperação desses passivos deve ocorrer por meio da regeneração natural, o que inclui o manejo de clareiras para plantio e condução da regeneração natural, conforme previsto no Artigo 44 da Instrução Normativa 05 da SEMAS/PA, publicada em 11 de setembro de 2015. O conhecimento da regeneração, recrutamento, crescimento e mortalidade de indivíduos de espécies comerciais em florestas tropicais manejadas é crucial para sua conservação e previsão sobre a produção futura destas florestas. Essas informações dão ao gerente florestal uma imagem muito mais clara sobre a capacidade de recuperação de uma

determinada floresta manejada. Isso, portanto, permite ao gerente tomar decisões precisas sobre a melhor forma de manejar sua floresta.

A abertura de clareiras no dossel causadas pela morte natural de árvores são distúrbios comuns de pequena escala e desempenham papel importante na dinâmica da floresta (WHITMORE, 1989). No entanto, clareiras artificiais causadas pelo corte de espécies arbóreas em florestas tropicais têm impactos maiores do que aqueles naturais. As alterações no ambiente florestal causadas pela exploração madeireira podem modificar a trajetória da regeneração natural, crescimento, mortalidade e recrutamento de indivíduos que compõem as comunidades destas florestas (SCHWARTZ et al., 2014; DE AVILA et al., 2017, DIONISIO et al., 2018).

A exploração florestal com técnicas de EIR é baseada em planejamento de operações e treinamento dos recursos humanos. Na EIR deve-se: a) minimizar danos ambientais, conservando serviços ambientais e potencial de exploração futura e; b) reduzir custos operacionais da exploração, aumentando a eficácia do trabalho, e c) reduzir desperdícios. As florestas sob EIR geralmente apresentam uma dinâmica mais alta do que a dinâmica observada em florestas intocadas, devido a operações de exploração como construção de infraestrutura (estradas, trilhas de arraste e pátios de estocagem), corte e arraste de árvores (YGUEL et al., 2019). Essas alterações na floresta muitas vezes promovem melhores condições para o crescimento de espécies remanescentes baseadas na dinâmica da sucessão de indivíduos (COSTA et al., 2020).

Apesar dos significativos avanços obtidos nas formas de exploração madeireira em florestas tropicais primárias, o que inclui a aplicação de EIR, não há sinais de sustentabilidade do sistema silvicultural no longo prazo (SIST e FERREIRA, 2007; VALLE et al., 2007). O que se tem demonstrado por meio de simulações e monitoramento de longo prazo, é que a maior parte das espécies exploradas não é capaz de recuperar os volumes explorados dentro de um ciclo (25-35 anos) completo de corte (REIS et al., 2010). Assim, as florestas tropicais primárias não são capazes de apresentar o mesmo rendimento de madeira no segundo ciclo de corte. Desta forma, a colheita florestal por meio de EIR em florestas tropicais primárias ainda precisa evoluir muito para alcançar uma exploração cíclica sustentável de madeira.

Como possíveis medidas para contornar o problema estão: a) mudança de mercados para espécies de baixa densidade de madeira (REIS et al., 2010); b) redução de volume colhido ou então alongamento do ciclo de corte (HAWTHORNE et al., 2012); c) redução do diâmetro mínimo de corte (DMC) para algumas espécies (SIVIERO et al., 2020b) e d) aplicação de tratamentos silviculturais pós-colheita (DE GRAAF et al., 1999; DAUBER et al., 2005; CARVALHO et al., 2008). As práticas silviculturais pós-colheita para aumentar as taxas de

crescimento das espécies arbóreas comerciais, o plantio de espécies de valor econômico (SCHWARTZ et al., 2017b) e o aproveitamento do volume perdido por mortalidade tornam-se fundamentais para equilibrar as perdas de madeira em florestas tropicais devido às colheitas (DIONISIO et al., 2018).

Os diferentes tratamentos silviculturais aplicados após a colheita, favorecem o estabelecimento de regeneração natural e maior produção em florestas tropicais (AZEVEDO et al., 2008; DE AVILA et al., 2017; SCHWARTZ et al., 2017a, 2017b; GOMES et al., 2019). Entre os tratamentos silviculturais pós-colheita estão aqueles que se concentram em garantir a reposição de indivíduos de espécies comerciais para os próximos ciclos de corte via plantio ou condução de regeneração natural. O plantio de espécies comerciais e a condução da regeneração natural encontram ambientes propícios para serem bem-sucedidos dentro das clareiras abertas pela queda de árvores colhidas (KUKKONEN et al., 2008). As técnicas de plantio e condução da regeneração natural vêm se desenvolvendo em diferentes florestas tropicais no mundo, como na Amazônia Oriental (LOPES et al., 2008; SCHWARTZ et al., 2013), Amazônia Ocidental (D'OLIVEIRA E RIBAS, 2011; KARSTEN et al., 2013) e Oeste da África (DOUCET et al., 2009). Assim, estes promissores tratamentos silviculturais têm merecido mais atenção em pesquisas científicas e desenvolvimento tecnológico.

Nesta tese serão abordados os efeitos a médio prazo de tratamentos silviculturais em clareiras oriundas da exploração de impacto reduzido, tanto de árvores comerciais plantadas quanto das de regeneração natural sob diferentes tratamentos, utilizando-se de técnicas que visam aumentar a densidade de espécies exploradas, de espécies com baixa densidade e/ou visando produção madeireira; que mantenham a variabilidade genética através das mudas e/ou sementes nativas; e que ao mesmo tempo conservem *in situ* as espécies podendo ainda ter utilização das mesmas nos próximos ciclos de corte. Conforme regimento geral da UFRA, que instrui escrita nas normas da revista a que serão submetidos, os capítulos/artigos estão no formato da revista *Forest Ecology and Management*.

OBJETIVOS

Geral

Avaliar o comportamento de espécies comerciais em clareiras de exploração de impacto reduzido.

Específicos

Capítulo 1:

Avaliar o efeito a médio prazo dos tratamentos silviculturais pós-colheita sobre a mortalidade, o crescimento e a estrutura de espécies comerciais de árvores, tanto plantadas quanto as de regeneração natural, presentes em clareiras de exploração de impacto reduzido.

Capítulo 2:

Avaliar o efeito a médio prazo dos tratamentos silviculturais pós-colheita em clareiras de diferentes tamanhos, com uso de árvores da espécie *Dinizia excelsa* plantadas e provenientes da regeneração natural sob diferentes níveis de sombreamento.

Capítulo 3:

Avaliar o efeito a médio prazo dos tratamentos silviculturais pós-colheita em clareiras utilizando árvores da espécie *Tachigali glauca* plantadas e provenientes da regeneração natural sob diferentes níveis de sombreamento.

QUESTÕES E HIPÓTESES

Capítulo 1:

Questão: Os tratamentos silviculturais pós-colheita influenciam na sobrevivência e crescimento de espécies comerciais madeireiras?

Hipótese nula: Os tratamentos silviculturais pós-colheita não influenciam na sobrevivência e crescimento de espécies comerciais em clareiras da exploração de impacto reduzido.

Hipótese alternativa: Os tratamentos silviculturais pós-colheita aumentam a sobrevivência e crescimento de espécies comerciais em clareiras da exploração de impacto reduzido.

Capítulo 2:

Questão: Qual efeito a médio prazo de tratamentos silviculturais em indivíduos de *Dinizia excelsa* em termos de mortalidade e crescimento?

Hipótese nula: Os tratamentos silviculturais pós-colheita não influenciam na sobrevivência e crescimento de *D. excelsa* em clareiras da exploração de impacto reduzido.

Hipótese alternativa: Árvores de *D. excelsa* não sofrem influência dos tamanhos de clareira, mas terão influência dos níveis de sombreamento, havendo alto crescimento no menor nível de sombreamento, tanto de indivíduos plantados quanto aqueles de regeneração natural.

Capítulo 3:

Questão: A condução da regeneração natural e enriquecimento de *Tachigali glauca* em clareiras mais a sua condução diminui as taxas de mortalidade e aumenta as de incremento?

Hipótese nula: Os tratamentos silviculturais pós-colheita não influenciam na sobrevivência e crescimento de *T. glauca* em clareiras da exploração de impacto reduzido.

Hipótese alternativa: Árvores de *T. glauca* terão influência dos níveis de sombreamento, havendo alto crescimento no menor nível de sombreamento, tanto de indivíduos plantados quanto aqueles de regeneração natural.

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1 **Post-harvesting silvicultural treatments in canopy logging gaps: medium-term responses**
 2 **of commercial tree species under tending and enrichment planting¹**

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8

9 **Abstract**

10 Technical and scientific information on medium-term effects of post-logging silvicultural
 11 interventions on the recovery of harvestable growing stocks are hardly available. To mitigate
 12 uncertainties about these effects our study aimed to answer the following question: “What is
 13 the medium-term effect of post-harvest silvicultural treatments on mortality, growth, and the
 14 structure of commercial tree species in canopy logging gaps under tending and enrichment
 15 planting?” we study individuals planted and naturally regenerated in 72 logging gaps opened
 16 tree felling during reduced-impact logging among different silvicultural treatments: (1) natural
 17 regeneration tending (TNER); (2) enrichment planting in logging gaps (EP1); (3) enrichment
 18 planting in logging gaps previously cleaned of harvesting residuals (EP2). Mortality increased
 19 through time, EP1 presented the highest mortality rates of all treatments in the first, sixth and
 20 11th year, TNER had the lowest at the same period. TNER and EP2 presented the highest basal
 21 area and EP2 the highest periodic annual increment. Effects of the silvicultural treatment TNER
 22 were positive, since it presented the highest survival and a high mean basal area of the initial
 23 trees. The medium-term effects of silvicultural treatments applied over individuals of
 24 commercial trees in logging gaps indicate higher survival and growth that reflected in the
 25 structure of treated individuals when compared to standard procedures of reduced impact-
 26 logging, these results bring positive outcomes to reach more sustainable future cutting cycles
 27 in the Brazilian Amazon and other tropical forests worldwide.

28 **Keywords:** assisted densification; conservation and timber production; sustainable forest
 29 management; tropical forest

30

31 **1. Introduction**

32 The fate of tropical forests has been guided by anthropic activities (Lewis and Maslin,

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33 2015; Putz et al., 2012). Species richness simplification in tropical environments has been
34 observed where species have been lost due to a combination of rapid climate change, natural
35 populations isolation in fragmented landscapes, competition against invasive species, and the
36 impact of increasing disturbances due to land use changes (Lewis et al., 2015; Lewis and
37 Maslin, 2015). To conciliate the conservation of forest biodiversity as well as other ecosystem
38 services with economic interests is a big challenge for nations, especially those developing
39 countries with large tropical forests (Chaudhary et al., 2016).

40 Reduced-impact logging (RIL) has been applied in tropical forests as an important tool
41 to mitigate destructive impacts of conventional logging in order to conserve biodiversity and
42 other ecosystem services as well to reduce deforestation rates (Putz et al., 2012; Schwartz et al.,
43 2012). Although its several benefits, other studies have shown that tree mortality increases after
44 harvesting following all RIL requirements (Bladon et al., 2008; Hautala and Vanha-Majamaa,
45 2006; Lavoie et al., 2012). Even after the application of RIL, mortality rates increase in response
46 to disturbance effects that can remain up to 11 years after harvesting (Dionisio et al., 2017).

47 Besides the increased mortality, simulations of future scenarios indicate that the current
48 technological advances in tropical forest management do not guarantee the same volume yields
49 of wood for further harvesting cycles (Avila et al., 2017; Dauber et al., 2005; Dionisio et al.,
50 2017; Hawthorne et al., 2012; Putz et al., 2012; Sist and Ferreira, 2007; Valle et al., 2007). The
51 scarcity of natural regeneration of commercial species in harvested tropical forests (Park et al.,
52 2005; Schwartz et al., 2017a; Van Rheenen et al., 2004), the large amount of timber volume
53 removal per harvest cycle, and the volume of wood lost due to increased post-harvest mortality
54 are crucial factors to prevent successful future harvesting cycles. Evidences show that the
55 current cutting cycle of 25-35 years employed in the Brazilian Amazonian forests with a
56 maximum harvesting volume of $30 \text{ m}^3 \text{ ha}^{-1}$ (SEMAS, 2015) cannot be sufficient to allow the
57 recovery of harvested timber volumes (Avila et al., 2017; Dionisio et al., 2018, 2017; Schwartz

58 et al., 2013; Sist and Ferreira, 2007). Non-logged commercial species, however, could be
59 harvested in further cutting cycles to ensure the same forest yields.

60 Some of the possible alternatives to mitigate this issue includes: a) to change the set of
61 harvested species to lighter wood species (Reis et al., 2010); and b) to apply post-harvesting
62 silvicultural treatments (Dauber et al., 2005) such as tending the natural regeneration and
63 planting commercial species in canopy gaps created by tree felling during harvesting operations
64 (Schwartz et al., 2017b). Silvicultural treatments increase timber production in tropical forests
65 and favor the establishment of natural regeneration and the growth of seedlings potentially able
66 to replace harvested individuals (Avila et al., 2017; Inada et al., 2017; Schwartz et al., 2017b,
67 2017a, 2013; Vieira et al., 2018; Villegas et al., 2009), but there is a need to assess the costs of
68 silvicultural treatments.

69 Technical and scientific information on medium-term effects of post-logging silvicultural
70 interventions on the recovery of harvestable growing stocks are hardly available (Petrokofsky
71 et al., 2015) and are empirical or based simply on simulations. In this sense, it becomes
72 necessary a non-empirical silvicultural system for the sustainable management of tropical forest
73 resources (Jardim, 2015). To mitigate uncertainties about the effects of post-harvest
74 silvicultural treatments (Gomes et al., 2019; Lopes et al., 2008; Schwartz et al., 2013; Souza et
75 al., 2015; Taffarel et al., 2014; Vieira et al., 2018) our study aimed to respond the following
76 questions: What is the medium-term effect of post-harvest silvicultural treatments on survival,
77 growth, and the structure of commercial tree species in canopy logging gaps under tending and
78 enrichment planting?

79

80 **2. Materials and methods**

81 **2.1 Study area and sampling design**

82 Data were obtained from a field experiment carried out in the forest management area of
83 the forestry company Jari Florestal SA under the project ‘Logging Gaps Management’
84 coordinated by Embrapa Eastern Amazon in cooperation with Jari Florestal. The study area is
85 located in the Jari valley, Almeirim municipality (1° 9' S, 52° 38' W), Pará state, Brazil. Average
86 annual precipitation is 2200 mm and the annual average temperature is 26 °C. The vegetation
87 is mainly ombrophilous dense forest over yellow latossols (Azevedo, 2006).

88 Jari Florestal has a total area of 545,535 ha under forest management where harvesting
89 operations follow RIL techniques. A total of 2,997 seedlings and saplings naturally present or
90 planted in 72 gaps created by tree felling due to RIL operations were assigned to assess
91 treatments as follows: (1) standard procedures of RIL (SRIL), (2) tending of the naturally
92 established regeneration (TNER), (3) enrichment planting 1 (EP1), and (4) enrichment planting
93 2 (EP2).

94 The experiment was established in 2006 and 2007 in the logging compartments harvested
95 in 2004 and 2006. In SRIL, which served as control, marked individuals were only monitored,
96 with no additional silvicultural treatments, according to the current forest management
97 regulations for forest monitoring in the Brazilian Amazon. In the other three treatments,
98 silvicultural procedures were applied in addition to all steps required to employ RIL.

99 Tending consisted in the liberation of target individuals against competing tree species
100 and lianas, it was applied over seedlings and saplings of commercial tree species naturally
101 established (TNER) and planted (EP1 and EP2) in all measurement years. In the enrichment
102 planting treatments, all seedlings were planted in a spacing of 2.5 m × 2.5 m using commercial
103 tree species of different ecological groups. EP1 and EP2 differed in terms of planted species,
104 logging compartments, gap ages and removal of logging residuals. EP1 was established in 2-
105 year gaps where no logging residual was removed while EP2 was established in 1-year gaps

106 with complete logging residual removal for further energy production by the forestry company
 107 (Table 1).

108

109 **Table 1.** Initial conditions (2006-2007) of the four post-harvesting silvicultural treatments
 110 tested (mean \pm SD) for seedlings and saplings naturally present or planted in logging gaps in
 111 Jari Florestal, Brazil (adapted from Schwartz et al., 2016).

Variable	Standard RIL (SRIL)	Tending of natural regeneration (TNER)	Enrichment planting 1 (EP1)	Enrichment planting 2 (EP2)
Number of seedlings and saplings	436	396	1520	645
Density (number of individuals / m ²)(mean \pm SD)	0.070 \pm 0.017	0.056 \pm 0.015	0.110 \pm 0.035	0.109 \pm 0.020
Number of species	34	39	10	5
Number of logging gaps	15	15	34	8
Size of logging gaps (m ²) (mean \pm SD)	395.1 \pm 86.3	478.4 \pm 200.4	418.7 \pm 97.6	754.9 \pm 237.7
Age of logging gaps at the beginning of the experiment (years)	2	2	2	1
Logging compartment	2004	2004	2004*	2006**
Logging residuals removed from gaps	No	No	No	Yes
Silvicultural treatment	No treatment	Tending the natural regeneration	Enrichment planting and tending	Enrichment planting and tending

112 * Species planted in 2006

113 ** Species planted in 2007

114

115 The logging gaps used in this experiment were set in terms of size, according to the
 116 classification of Jardim et al. (2007) in: 27 small size gaps (200–400 m²), 30 medium size gaps
 117 (401-600 m²), and 11 large size gaps (> 600 m²). EP2 has the largest logging gaps with a
 118 significant statistical difference of the other three treatments, whose did not differ among them

119 (SRIL, TNER, and EP1) in size. The tending silvicultural treatments on TNER, EP1, and EP2
 120 were applied annually from 2006 to 2010, 2012, 2017, and 2018. In 2018 the treatments SRIL,
 121 TNER, EP1, and EP2 presented 203, 248, 461, and 276 trees, respectively, which remained
 122 alive since the experiment beginning (henceforth “Initials”). At the 2017 measurement, new
 123 individuals ≥ 300 cm in height that grew naturally inside the monitored logging gaps
 124 (henceforth “Complementary”) were included and monitored in the experiment. All species of
 125 the study are commercial but not necessarily were harvested by the forestry company. The total
 126 number of new individuals included in SRIL, TNER, EP1, and EP2 were 51, 44, 129, and 20,
 127 respectively (Table 2).

128

129 **Table 2.** Current conditions (2017-2018) of the four post-harvesting silvicultural treatments
 130 tested for initial and complementary individuals in logging gaps in Jari Florestal, Brazil.

Variable	Standard RIL (SRIL)	Tending of natural regeneration (TNER)	Enrichment planting 1 (EP1)	Enrichment planting 2 (EP2)
Number of alive trees \geq 300 cm in height since the experiment beginning	203	248	462	274
Number of complementary trees, included in 2017	51	44	129	20
Total number of trees in the experiment	254	292	591	294
Number of species included in 2017	0	0	15	8
Current total number of species	29	34	24	13
Current number of logging gaps	15	13	32	8
Total area of logging gaps (m ²)	5,926.61	6,587,13	13,325.85	6,038.93

131

132 2.2 Data analysis

133 Only individuals with height ≥ 300 cm were used in the analyses carried out in this study.
 134 This means that every planted or naturally regenerated seedling that did not reach minimum
 135 height of 300 cm by 2010 was not considered in this study. Mortality, diameter at breast height
 136 (DBH) at 1.3 m from the soil, periodic annual increment (PAI), basal area, and diameter classes
 137 were calculated based on the measurements of the years 2017 for SRIL, TNER, and EP1 and
 138 2018 for EP2, so in this way all treatments were analyzed under the same measurement ages.

139 Mortality rates were calculated with logging gaps as the sampling units. The annualized
 140 mortality rates were calculated using the formula “ $m = 1 - (N_{t2} / N_{t1})^{(1/t)}$ ”, where N_{t1} = Number
 141 of live trees in the first sampling, N_{t2} = number of trees that survived until the second sampling
 142 and t = years between first and second sampling (Sheil et al., 1995). Mortality rates did not
 143 follow a normal distribution, thus they were analyzed through a general linear model (GLM)
 144 repeated measures ANOVA, with time and treatment as factors and mortality rates as the
 145 dependent variable and compared by post-hoc Tukey’s pairwise test.

146 The logging gap area was calculated by the ellipse formula. Logging gap was also the
 147 sampling unit to calculate basal area, which was the sum of tree cross sections of each logging
 148 gap. Each tree cross section was obtained with the formula $g = \pi * (DBH/2)^2$ in square meters
 149 per hectare, where DBH = diameter at breast height. Basal area for the initial, complementary,
 150 and the sum of initial and complementary individuals (henceforth “Overall”) in each logging
 151 gap per treatment were calculated. Once applied the Shapiro-Wilk test, the data on both initial,
 152 complementary, and overall trees did not present a normal distribution, which required a Box-
 153 Cox transformation, before running ANOVA and the post-hoc Tukey’s pairwise test.

154 Periodic annual diameter (PAI) measured from DBH of each individual was calculated
 155 using the formula $PAI = ((DBH_{t2} - DBH_{t1})/n)*365$, where DBH_{t1} = individual’s diameter at
 156 the initial sampling, DBH_{t2} = individual’s diameter at the final sampling, and n = days between

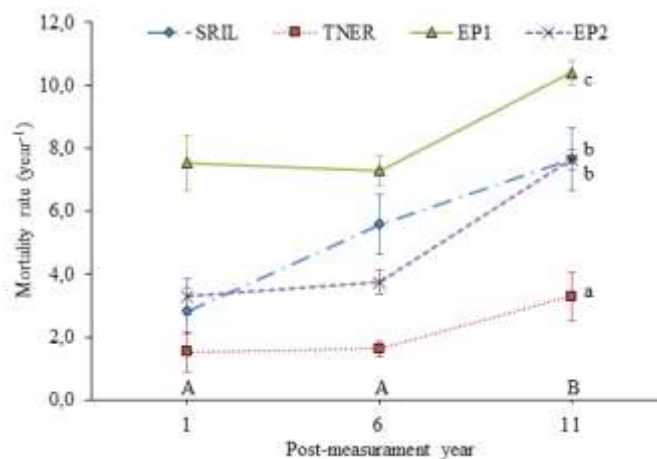
157 first and second sampling. Individual PAI was the sampling unit, ANOVA ($p < 0.05$
 158 significance level) was performed and compared by the post-hoc Tukey's pairwise test. All
 159 analyses were performed using the R version 3.0.2 (2016).

160 The number of individuals (initial, complementary, and overall trees), in percentages,
 161 were set in five diameter classes ranging 5 cm in DBH from 0 up to 25 cm. In addition, the
 162 basal area was also set in each of these five diameter classes.

163

164 3. Results

165 Mortality increased through time. In the first, sixth and 11th year, EP1 presented the
 166 highest mortality rates of all treatments, 7.5, 7.3, and 10.4% year⁻¹, respectively. TNER had the
 167 lowest mortality rate observed at the same period, with 1.5, 1.6, and 3.3% year⁻¹, respectively.
 168 There was a significant statistical difference among treatments (repeated measures ANOVA, p
 169 < 0.001) and time (repeated measures ANOVA, $p < 0.001$), with no significant interaction
 170 between treatment and time (repeated measures ANOVA, $p = 0.063$). SRIL and EP2 treatments
 171 presented statistical similarity ($p = 0.93$ significance level; Figure 1).

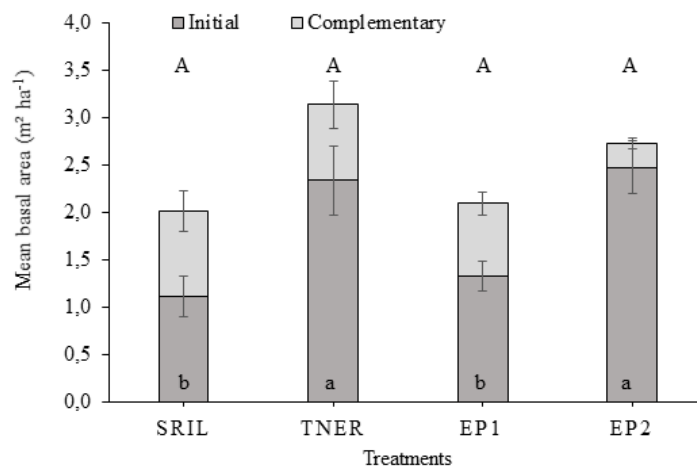


172

173 **Figure 1.** Mean (\pm SE) of mortality rates of standard procedures of reduced-impact logging
 174 (SRIL), tending of the naturally established regeneration (TNER), enrichment planting 1 (EP1),
 175 and enrichment planting 2 (EP2) treatments by one, six, and 11 years in logging gaps in the
 176 managed forests of Jari company, Pará state, Brazil. Lowercase letters are treatment
 177 comparisons and uppercase letters are time comparisons in the GLM repeated measures
 178 ANOVA and post-hoc Tukey's pairwise test.

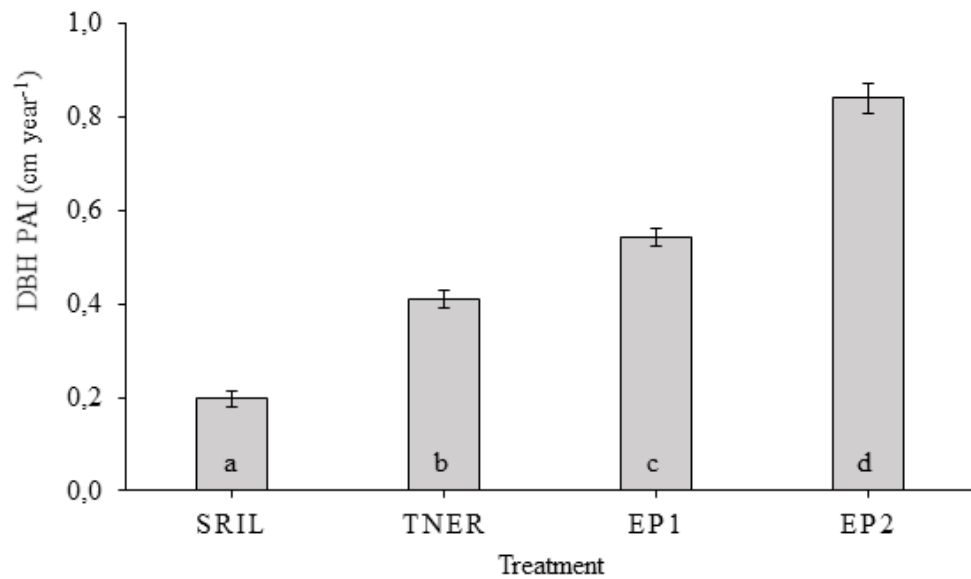
179

180 In terms of basal area of the initial trees 11 years after the experiment started, EP2
 181 presented the highest value ($2.47 \text{ m}^2 \text{ ha}^{-1}$) and SRIL the lowest ($1.12 \text{ m}^2 \text{ ha}^{-1}$). However, for the
 182 complementary trees, the highest basal area was observed in SRIL ($0.90 \text{ m}^2 \text{ ha}^{-1}$) and the lowest
 183 in the EP2 treatment ($0.25 \text{ m}^2 \text{ ha}^{-1}$). The sum of initial and complementary basal area resulted
 184 in TNER ($3.14 \text{ m}^2 \text{ ha}^{-1}$) as having the highest basal area, and SRIL with the lowest ($2.01 \text{ m}^2 \text{ ha}^{-1}$)
 185 ¹; Figure 2). Comparisons among initial trees presented statistical significant difference ($p <$
 186 0.001) among treatments, except between TNER ($2.33 \text{ m}^2 \text{ ha}^{-1}$) versus EP2 ($2.47 \text{ m}^2 \text{ ha}^{-1}$) and
 187 SRIL ($1.19 \text{ m}^2 \text{ ha}^{-1}$) versus EP1 ($1.18 \text{ m}^2 \text{ ha}^{-1}$) that had no statistical differences. No statistical
 188 significant differences were found among treatments in the mean basal area of the
 189 complementary and overall trees (Figure 2).



190 **Figure 2.** Mean (\pm SE) of basal area ($\text{m}^2 \text{ ha}^{-1}$) of standard procedures of reduced-impact logging
 191 (SRIL), tending of the naturally established regeneration (TNER), enrichment planting 1 (EP1),
 192 and enrichment planting 2 (EP2) treatments from the initial trees, complementary and the
 193 overall individuals (sum of initial and complementary individuals) in logging gaps in the
 194 managed forests of Jari company, Pará state, Brazil. Lowercase letters are to comparisons
 195 between the initial mean basal areal of treatments and uppercase letter are to mean basal area
 196 of complementary and overall in the ANOVA and post-hoc Tukey's pairwise test.
 197
 198

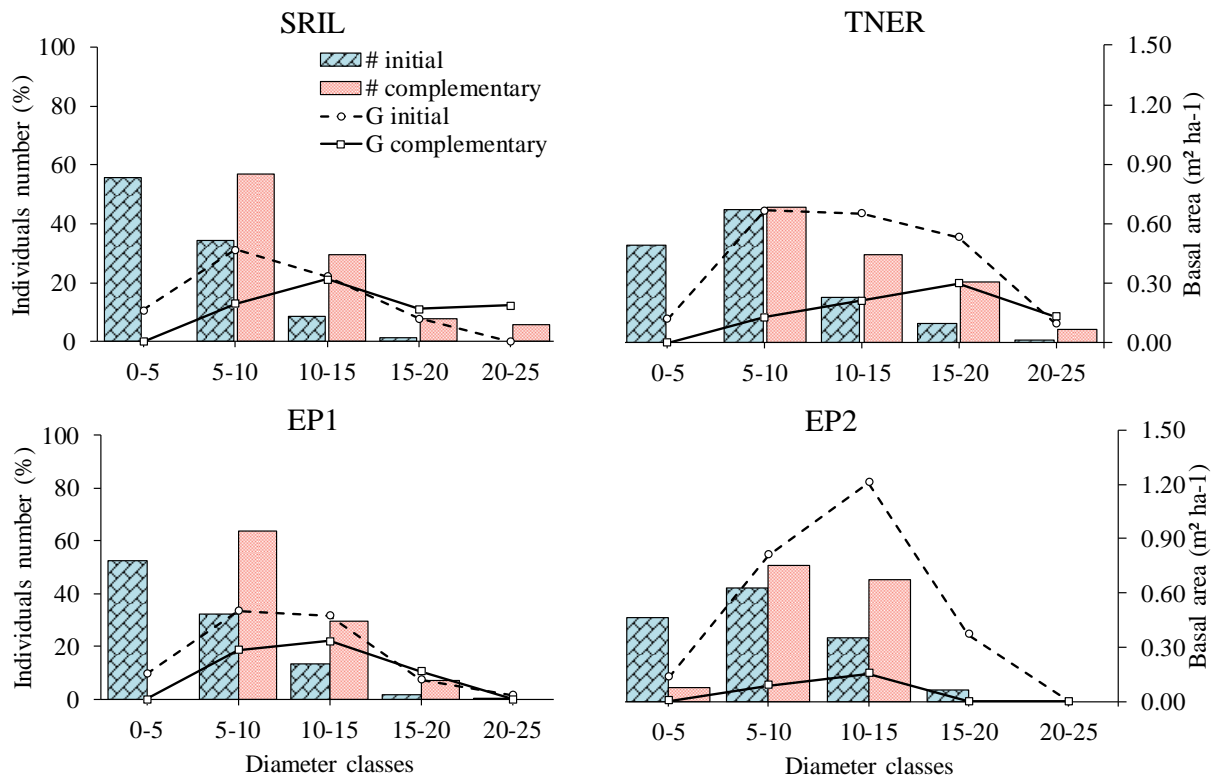
199 The lowest mean of PAI were observed in the natural regeneration treatments, where
 200 SRIL presented half ($0.20 \text{ cm year}^{-1}$) of the PAI mean observed in TNER ($0.41 \text{ cm year}^{-1}$). EP2
 201 presented the highest PAI mean ($0.84 \text{ cm year}^{-1}$), four times more than SRIL ($p < 0.05$).
 202 Statistical differences were found in all treatments ($p < 0.05$, Figure 3).



203 **Figure 3.** Mean (\pm SE) of periodic annual increment (PAI) of reduced-impact logging (SRIL),
 204 tending of the naturally established regeneration (TNER), enrichment planting 1 (EP1), and
 205 enrichment planting 2 (EP2) treatments in logging gaps in the managed forests of Jari company,
 206 Pará state, Brazil. Lowercase letters are treatment comparisons in ANOVA with post-hoc
 207 Tukey's pairwise test.
 208
 209

210 Over 11 years the 0-5 and 5-10 classes presented the highest percentage of the initial trees
 211 in all treatments. Meantime the 5-10 and 10-15 classes presented the highest percentage of the
 212 complementary trees. Among treatments, SRIL 0-5 cm diameter class presented the highest
 213 number of individuals. On the other hand, EP1 showed the highest concentration of individuals
 214 in class 5-10 cm of the complementary individuals. The basal area presented a paraboloid
 215 tendency for all treatments. However, the classes with the highest concentration of basal area
 216 are in the range of 5-15 cm (Figure 4).

217



218
 219 **Figure 4.** Bars representing percentage of individuals per diameter class (left side y-axis) and
 220 lines showing basal area ($\text{m}^2 \text{ha}^{-1}$) per diameter class (right y-axis) of the standard reduced-
 221 impact logging (SRIL), tending of the naturally established regeneration (TNER), enrichment
 222 planting 1 (EP1), and enrichment planting 2 (EP2) treatments for both initial and
 223 complementary trees in logging gaps in the managed forests of Jari company, Pará state, Brazil.
 224

225 4. Discussion

226 Effects of the silvicultural treatment TNER were positive, since it presented the highest
 227 survival and a high mean basal area of the trees. Furthermore, TNER presented twofold PAI
 228 value when compared to SRIL, which is the treatment that reflects the current procedures of
 229 reduced impact logging (RIL) legally adopted in the Brazilian Amazon. Tending treatment
 230 applied to natural regeneration and enrichment planting in logging gaps are financially
 231 profitable options for forest managers and investors (Schwartz et al., 2016) to minimize post-
 232 logging losses due to increased mortality observed in managed forests (Dionisio et al., 2017).
 233 These silvicultural treatments also help to mitigate the non-recovery of species diversity after
 234 logging observed in selectively logged forests (Shima et al., 2018).

235 The highest TNER survival is probably due to the fact that sampled individuals were

236 already established before the silvicultural treatments had started. Besides this, tending begun
237 to be applied to avoid competition for light against other arboreal individuals and lianas
238 (Schwartz et al., 2016). In this study, TNER was shown to be effective to increase survival of
239 individuals when compared to the other treatments tested. Higher mortality in enrichment
240 planting, like EP1 and EP2, in relation to natural regeneration, like SRIL and TNER is probably
241 due to the rustification phase (Adenesky-Filho et al., 2017).

242 Although the rustification phase, EP2 and SRIL did not present significant statistical
243 difference in mortality, which can figure as a positive treatment effect of cleaning potential
244 competitors for light and nutrients in enrichment planting. The higher mortality of enrichment
245 planting treatments in relation to tending may be a result of the low quality many planted
246 seedlings. The availability of high quality seedlings of native commercial species is a serious
247 bottleneck for enrichment planting in the Brazilian Amazon. For the success of enrichment
248 planting in logging gaps, it is recommended a rigorous seedling production, with fertilizers, as
249 well as annual cleanings once seedlings planted in the field normally face strong competition
250 (Gomes et al., 2019).

251 The highest complementary mean basal area in SRIL can be explained by the absence of
252 any silvicultural treatment, which permitted new individuals of commercial species to get
253 established. The fact that there was no tending over the monitored individuals of commercial
254 species (initials) could have allowed competing individuals to over compete them inside
255 logging gaps (Avila et al., 2017; Jardim, 2015). Results on basal area found in this study
256 corroborate Inada et al., (2017), in a comparison among RIL + Line Planting/Slashing (LP/S)
257 of tree species versus conventional logging and versus RIL. After 10 years, stocks were
258 marginally higher in RIL + LP/S plots, with no statistical difference among treatments.

259 SRIL, TNER, and EP2 presented a similar total area of logging gaps (m^2), but the basal
260 area of the initial trees of TNER and EP2 were twice higher than SRIL. The tending treatment

261 benefits both natural regeneration and enrichment planting and may be influenced by gap sizes,
262 since in larger gaps more light illuminates the forest floor, stimulating tree growth (Vatraz et
263 al., 2016). In an experiment with *Cedrela odorata* planted in logging gaps, Vieira et al. (2018)
264 found after five years that tending silvicultural treatment provided higher growth in height of
265 seedlings planted in larger logging gaps (>600 m²).

266 The canopy height formed by trees that surround a logging gap is crucial for the success
267 of individuals of tree species under any silvicultural treatment. Foresters can also enlarge
268 logging gaps in order to improve success of survival and growth rates of planted or tended
269 individuals of tree species. For example, two gaps with the same area but different surrounding
270 canopy heights will have different sunlight incidence, which will necessarily have effects on
271 the individual performances. *S. parahyba* planted in small gaps (200 m²) can thrive since the
272 surrounding canopy height is not so tall. And this is the case of Schwartz et al. (2017b) where
273 *S. parahyba* was managed in logging gaps surrounded by a short canopy height.

274 Low densities or absence of the regeneration of commercial species found in three
275 managed forests sampled by Schwartz et al., (2017a) suggest that commercial species are
276 having poor post-harvesting capacity to regenerate in the Eastern Amazon. The enrichment
277 planting in gaps is defined by Schwartz and Lopes (2015) as a type of assisted densification.
278 This means that tree species have their densities increased in their own natural habitats,
279 which can work both for conservation and timber production, as it can increase artificial
280 density up to 60 times compared to the natural densities of many commercial species.
281 Mahogany is an example of rare commercial species that can have its natural densities
282 increased (Lopes et al., 2008).

283 EP1 started with 10 species in 2006 and reached 24 in 2017, even though one of the
284 planted species had 100% of mortality rate in the period. The high mortality of EP1 trees and
285 the application of cleaning opened space for new individuals of fast-growing species. This

286 explains the entrance of the largest number of commercial species and individuals, mainly in
287 the higher diameter classes, of EP1. This outcome of silvicultural treatment application can
288 mitigate the sequential depletion of species (Putz et al., 2012) through the enrichment planting
289 and augment of the regeneration with the entrance of new species and stock of the most
290 commercial species. Most of the individuals present in higher diameter classes will probably
291 reach the minimum cutting diameter faster than initial trees.

292 EP2 was the most suitable treatment for the specific purpose of timber production and/or
293 species conservation. Furthermore, the improvement in light conditions caused by silvicultural
294 treatments improved the increase of both growth and stock of commercial species (Inada et al.,
295 2017). Successful enrichment planting in logging gaps confirms the results found in other
296 experiments worldwide with enrichment planting in gaps (Doucet et al., 2009; Gomes et al.,
297 2019, 2010; Lopes et al., 2008; Quédraogo et al., 2014; Schwartz et al., 2013; Taffarel et al.,
298 2014; Vieira et al., 2018). These studies help to reinforce the efficiency of enrichment planting
299 in logging gaps, which comes as a viable silvicultural alternative for managing tropical forests.

300 Enrichment planting in logging gaps can also work as an active germplasm bank to
301 maintaining the genetic diversity of rare or endangered species (Lopes et al., 2008) and may
302 have potential as seed/seedling orchard. Besides the use in old-growth managed forests, as the
303 forest management area of the forestry company Jari Florestal, such silvicultural treatments can
304 also be applied to recover degraded forests or to improve secondary forests, which are
305 commonly found in the arc of deforestation (Schwartz and Lopes, 2015), a region that
306 concentrates most of the deforestation in the Brazilian Amazon. To have this achievement, it is
307 necessary the development of new public policies to improve the current forest management
308 regulations in the Brazilian Amazon.

309 Silvicultural treatments also showed positive effects on the structure of individuals
310 in logging gaps. Based on these results, it is hypothesized that the silvicultural treatments

311 will shorten the time required to recover losses caused by RIL. Post-harvest silvicultural
312 treatments could help to mitigate the delay in several tropical forests to recover harvested
313 stocks (Shima et al., 2018).

314

315 **5. Conclusion**

316 The medium-term effects of silvicultural treatments applied over individuals of
317 commercial trees in logging gaps indicate higher survival and growth that reflected in the
318 structure of treated individuals when compared to standard procedures of reduced impact-
319 logging. These results bring positive outcomes to reach more sustainable future cutting cycles
320 in the Brazilian Amazon.

321

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328

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517

1 **Silviculture of high value timber species in the Amazon: managing *Dinizia excelsa* in** 2 **logging gaps**

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10

11 **Abstract**

12 *Dinizia excelsa*, one of the most important species traded in the Brazilian Amazon due to the
13 high commercial value of its timber, was analyzed through the medium-term effect of post-
14 harvesting silvicultural treatments in canopy logging gaps of different sizes under tending and
15 enrichment planting under different crown exposure. We hypothesized that individual effects
16 will have low or neither influence from gap sizes, but will have influence on the crown exposure
17 class. A total of 244 seedlings and saplings of *Dinizia excelsa* naturally present or planted in 43
18 gaps were assigned to assess the species medium-term responses to the following treatments:
19 (1) standard procedures of RIL (SRIL) or control, (2) tending of the naturally established
20 regeneration (TNER) and (3) enrichment planting (EP). There is better survival and growth in
21 enrichment planting (EP) and tending of natural regeneration (TNER) in canopy logging gaps
22 when compared to SRIL, which is the treatment that reflects the current procedures of reduced
23 impact logging (RIL) legally adopted in the Brazilian Amazon. There is no difference in the
24 growth of individuals according to the gap sizes, but according to their crown exposure class.
25 Despite the high mortality of planted individuals (EP), there was a better performance in
26 growth, especially those individuals who received more sunlight. These results bring positive
27 outcomes to reach more sustainable future cutting cycles to *D. excelsa* specie in the Brazilian
28 Amazon.

29 **Keywords:** assisted densification; conservation and timber production; sustainable forest
30 management; tropical forest.

31 **1. Introduction**

32 *Dinizia excelsa* Ducke (family Fabaceae), a tree species with wide geographical
33 distribution in the Eastern Amazon, is now represented by the tallest alive tree in the whole
34 biome, measuring 88 m in height and 5.5 m in diameter (Gorgens et al., 2019). *Dinizia excelsa*
35 figures currently as one of the most important commercial species traded in the Brazilian

36 Amazon due to its high timber value (Lopes De Souza et al., 2014). According Secretary for
37 Environment and Sustainability of Pará state, Brazil, a total of 1,486,952.87 m³ of *D. excelsa*
38 was traded in a recent period of 10 years (01/January/2006 to 21/February/2016) with monetary
39 values of US\$ 52,501,435.82, which gives a mean price of US\$35.85 ± 6.42 SD (BRL/USD
40 exchange rate of 15/February/2020, SEMAS/PA, 2020).

41 Despite of the substantial harvested timber volumes in the Brazilian Amazon, *D. excelsa*
42 presents naturally low densities of individuals because of its light requirements, only found in
43 canopy gaps opened by forest disturbances (Cysneiros et al., 2018). Forest disturbances,
44 including those caused by logging, that create canopy gaps, can be used as an efficient way of
45 conserving low density or rare tree species (Gomes et al., 2019; Neves et al., 2019; Schwartz
46 and Lopes, 2015; Vieira et al., 2018). In this context, the application of silvicultural treatments
47 can also work as an effective procedure for a more sustainable economic use of these species
48 (Gomes et al., 2019; Schwartz et al., 2016).

49 Post-harvesting silvicultural treatments applied over individuals in canopy logging gaps,
50 as tending and enrichment planting, bring positive outcomes in order to reach more sustainable
51 future cutting cycles in tropical native forests (Doucet et al., 2009; Gomes et al., 2019, 2010;
52 Lopes et al., 2008; Neves et al., 2019; Quédraogo et al., 2014; Schwartz et al., 2013; Taffarel
53 et al., 2014; Vieira et al., 2018). Tending consists in the liberation of target individuals against
54 competing non-commercial tree species and lianas, providing better light availability to
55 improve survival and more rapid growth (Brokaw 1985, Brown & Whitmore 1992). Therefore,
56 these silvicultural treatments applied to *D. excelsa* can also contribute to increase the species'
57 natural low densities through assisted densification (Schwartz and Lopes, 2015) that can result
58 in increments on timber production ensuring the species conservation in further cutting cycles.

59 The tree species regeneration inside micro-environments as canopy gaps is higher in
60 large gaps than small gaps (Buajan et al., 2018). The canopy height formed by surrounding
61 trees of a logging gap is determinant variable, even more important than gap size, for the success
62 of tree species individuals under any silvicultural treatment. Schwartz et al. (2017) studying
63 showed the successful performance of the pioneer species *Schyzolobium parahyba* var.
64 *amazonica* planted in small managed logging gaps (200 m²) surrounded short canopy height.
65 So, treated individuals of commercial species in logging gaps can thrive since the surrounding
66 canopy height is not so tall, what optimizes sunlight incidence. One of the methods used to
67 categorize sunlight incidence over trees is the crown exposure class system (CEC) of Shenkin
68 et al. (2018), adapted of Clark and Clark (1992).

69 In the Brazilian Amazon, managers have invested little or even no resources in post-

70 harvest treatments in Sustainable Forest Management Plans (SFMP). Research with positive
71 outcomes, including positive cost-benefit analyses, of studies about post-harvesting
72 silvicultural treatments in canopy logging gaps (Gomes et al., 2019, 2010; Neves et al., 2019;
73 Schwartz et al., 2016, 2013; Vieira et al., 2018) with high value commercial species as *D.*
74 *excelsa*, can increase the interest of managers. Thus, the objective of this study was to analyze
75 the medium-term effect of post-harvesting silvicultural treatments in canopy logging gaps of
76 different sizes under tending and enrichment planting with *D. excelsa* individuals under
77 different crown exposure. We hypothesized that, after application of the silvicultural treatments
78 on trees in logging gaps, individual effects will have low or neither influence from gap sizes,
79 but will have influence on the crown exposure class, mainly with high growth in the highest
80 CEC, regardless gap sizes.

81

82 **2. Materials and methods**

83 2.1. *Study area*

84 Data were collected in a field experiment carried out in the forest management area of the
85 forestry company Jari Florestal SA under the project 'Logging Gaps Management' coordinated
86 by Embrapa Eastern Amazon in cooperation with Jari Florestal. The company has a total forest
87 management area of 545,535 ha and all harvesting operations follow RIL techniques. The study
88 area is located in the Jari valley, Almeirim municipality (1° 9' S, 52° 38' W), Pará state, Brazil.
89 Average annual precipitation is 2200 mm and the annual average temperature is 26 °C. The
90 vegetation is mainly ombrophilous dense forest or *terra firme* forest, where the most common
91 soils are yellow latossols (Azevedo, 2006).

92 2.2. Experimental design

93 A total of 244 seedlings and saplings of *Dinizia excelsa* naturally present or planted in 43
94 gaps created by tree felling under RIL were assigned to assess the species medium-term
95 responses to the following treatments: (1) standard procedures of RIL (SRIL) or control, (2)
96 tending of the naturally established regeneration (TNER) and (3) enrichment planting (EP). All
97 seedlings and saplings had 40-80 cm in height.

98 The experiment was established in 2006 in the logging compartments harvested in 2004.
99 In SRIL, individuals of commercial species were only monitored, with no additional
100 silvicultural treatments, following the current forest regulations for forest monitoring in the
101 Brazilian Amazon. In the other two treatments, silvicultural procedures were applied in addition
102 to all steps required for RIL. Tending consisted in the liberation of target individuals against

103 competing individuals of tree species and lianas. This treatment was applied over and around
 104 seedlings and saplings of commercial tree species naturally established (TNER) and planted
 105 (EP), annually from 2006 to 2010, 2012 and 2017. In the EP treatment all seedlings were
 106 transplanted from the annual production unit (APU) to the logging gaps in a planting spacing
 107 of 2.5 m × 2.5 m inside in 2-year-old logging gaps.

108 According to the pre-logging inventory (PL) done in 2003, as the standard procedures of
 109 RIL require, the annual production unit (APU) area of 7,600 ha presented 5,300 *D. excelsa*
 110 individuals (N). Pre-logging absolute density (ADe), dominance (ADo) and basal area (G) of
 111 *Dinizia excelsa* were calculated (Table 1). Forest characteristics as volume, species
 112 composition, diameter structure and diversity indexes of the study area (annual production unit)
 113 of the study area are available in Lopes De Souza et al. (2014).

114

115 Table 1: Measurement year, number of individuals (N), basal area (G), absolute density (ADe) and dominance
 116 (ADo) of *Dinizia excelsa* in the pre-logging and in the annual production unit of 2004 in the forest management
 117 are of Jari S.A., Eastern Amazon. PL (pre-logging), SRIL (standard reduced impact logging), TNER (tending of
 118 the naturally established regeneration), and EP (enrichment planting).

	PL*	SRIL	TNER	EP
Measurement year	2003	2006	2006	2006
Number of individuals in 2006	5300	57	45	142
Number of individuals in 2017	-	39	36	72
Area (ha)	7600	0.593	0.659	1.333
G (m ² ha ⁻¹)	0.49	0.06	0.14	0.26
ADe (ind ha ⁻¹)	0.7	65.8	54.65	54.78
ADo (m ² ha ⁻¹)	0.49	0.1	0.21	0.2

119 * The *D. excelsa* Pre-logging (PL) information is to the year 2003

120

121 In 2017 only individuals that reached ≥ 300 cm in height in the measurement of 2012
 122 were included in the analyses, so the total number of *D. excelsa* individuals was 20, 32 and 65
 123 for SRIL, TNER and EP, respectively.

124 Mortality rates were calculated using the formula “ $m = 1 - (Nt2 / Nt1)(1/t)$ ”, where $Nt1$
 125 = Number of live trees in the first sampling, $Nt2$ = number of trees that survived until the second
 126 sampling and t = years between first and second sampling (Sheil et al., 1995). The logging gap
 127 area was calculated by the ellipse formula. The basal area was the sum of tree cross sections of
 128 each treatment. Each tree cross section was obtained with the formula $g = \pi * (DBH/2)^2$ in
 129 square meters per hectare, where DBH = diameter at breast height. The number of individuals
 130 in percentages, were set in diameter class ranging 4 cm in DBH from 0 up to 16 cm. Percentage
 131 of individuals per diameter class (0-4 cm; 4-8 cm, 8-12 cm and 12-16 cm) distributed in gap

132 sizes, treatments and crown exposure classes, over 11 years in logging gaps in the managed
133 forests of Jari S.A.

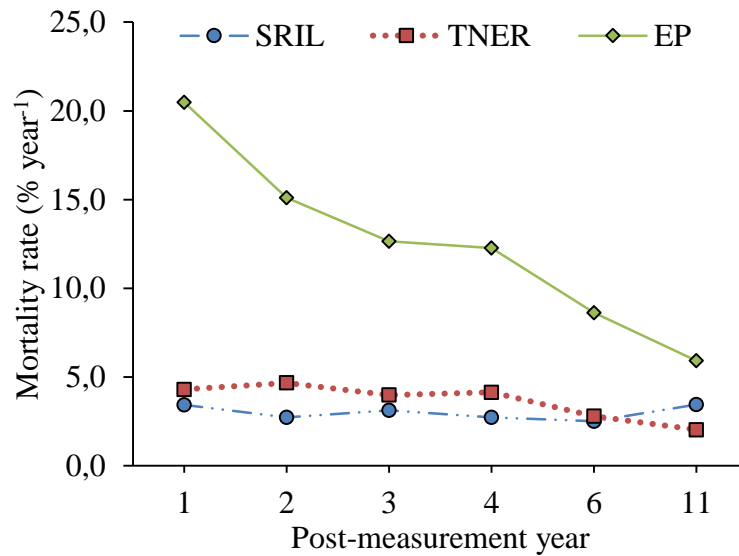
134 Periodic annual diameter (PAI) was obtained from diameter at breast height (DBH) at 1.3
135 m from the soil of each individual through the formula $PAI = ((DBHt2 - DBHt1)/n)*365$,
136 where DBHt1 = individual's diameter at the initial sampling, DBHt2 = individual's diameter at
137 the final sampling, and n = days between first and second sampling. The 117 individuals of *D.*
138 *excelsa*, which reached minimum height of 300 cm by 2012, were compared by treatment, gap
139 size (GS) and crown exposure class (CEC). SRIL, TNER and EP presented respectively 20, 32
140 and 65 individuals. To permit gap size comparisons, the individuals were set in: 59 individuals
141 in small size gaps (200–400 m²) and 58 individuals in large size gaps (> 400 m²). The crown
142 exposure of all seedlings and saplings were classified since the beginner of the study before and
143 after the application or not of tending silvicultural treatment, the classification was divided in
144 classes according to Clark & Clark (1992) in: 1 (no direct light, 51 individuals), 2 (some lateral
145 light, 39 individuals) and 3 (10–90% overhead light, 27 individuals).

146 Once applied the Shapiro-Wilk test on PAI by treatment, gap size and crown exposure
147 class, the data did not present a residual normal distribution ($p_{normal} < 0.01$). This required a
148 Box-Cox transformation, before running ANOVA and the post-hoc Tukey's pairwise test to
149 gap sizes, treatments and crown exposure class. Besides these analyzes, a two-way ANOVA
150 and the post-hoc Tukey's pairwise were tested too to the treatments versus crown exposure
151 class. All analyses were performed using the R version 3.0.2 (2016).

152

153 **3. Results**

154 Mortality rates of the natural regeneration treatments remained constant, except EP in the
155 initial years of the experiment. Over 11 years EP presented the highest mortality rate (6% year⁻¹)
156 while TNER had the lowest rate (2% year⁻¹, Figure 1).

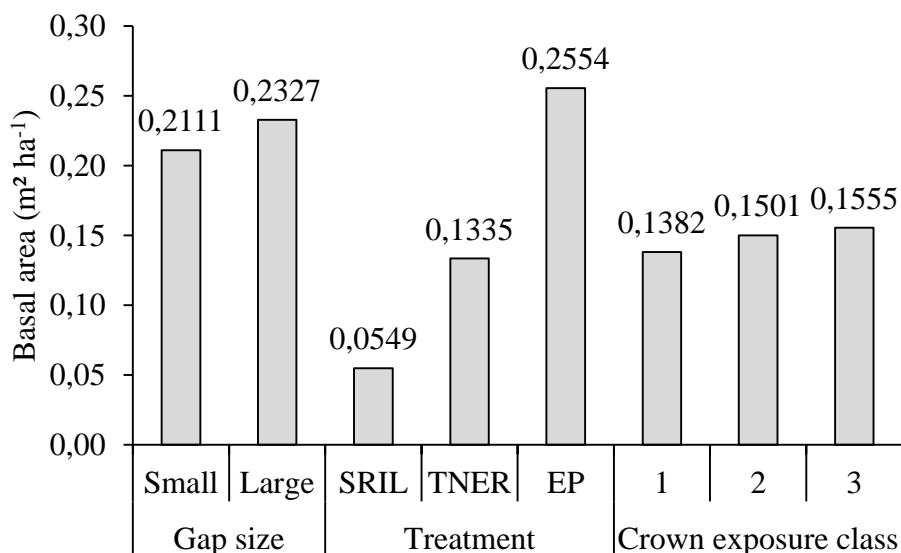


157

158 Figure 1: Mortality rates of standard procedures of reduced-impact logging (SRIL), tending of the naturally
 159 established regeneration (TNER) and enrichment planting (EP) treatments over 11 years in logging gaps in the
 160 managed forests of Jari S.A., Eastern Amazon, Brazil.

161

162 The highest basal areas were observed in large gaps, EP treatment and crown exposure
 163 class (CEC) 3. The lowest basal area was observed in small gap sizes, SRIL treatment and to
 164 CEC 1 (Figure 2).



165

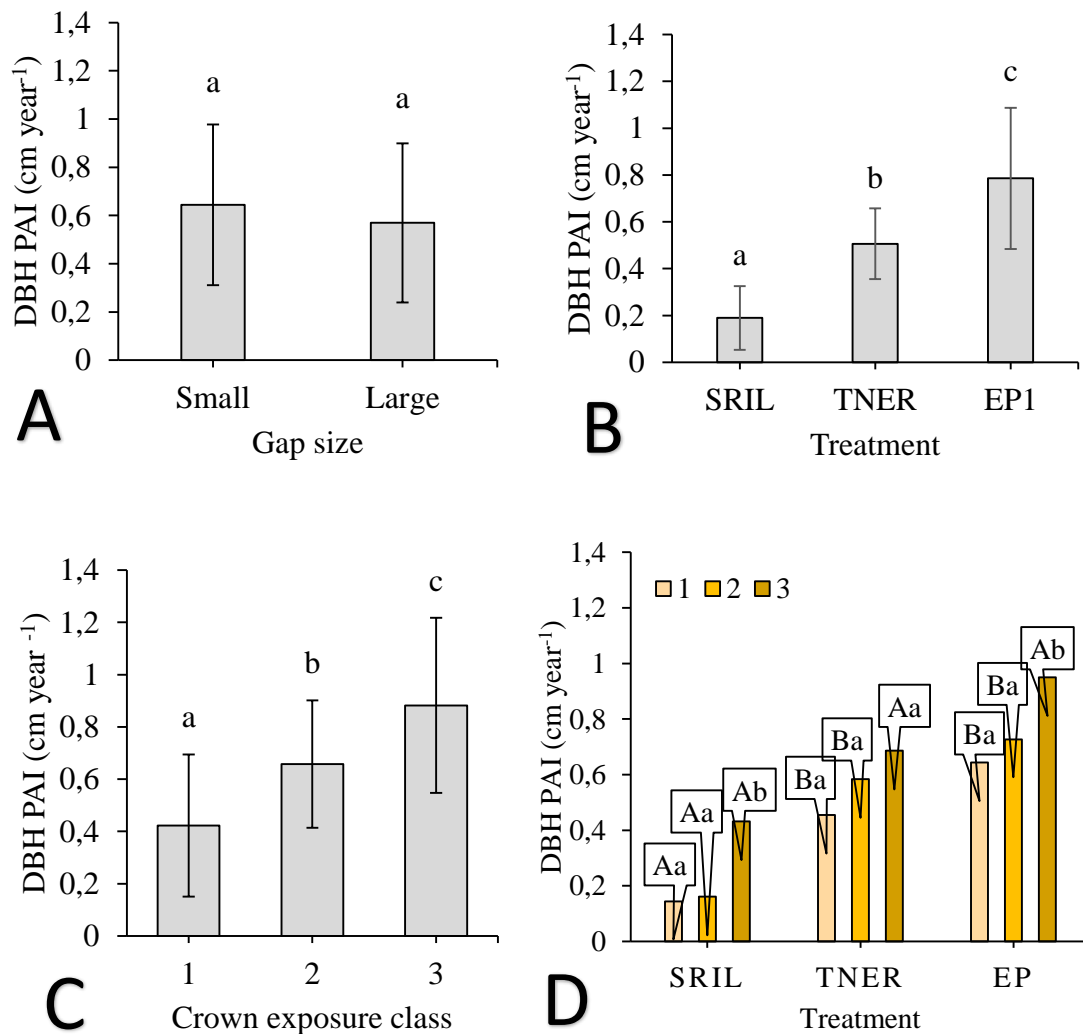
166 Figure 2: Basal area (m² ha⁻¹) of standard procedures of reduced-impact logging (SRIL), tending of the naturally
 167 established regeneration (TNER) and enrichment planting (EP) treatments over 11 years in logging gaps in the
 168 managed forests of Jari S.A., Eastern Amazon, Brazil.

169

170 Taking into consideration only the PAI effect on gap sizes, there was no statistical
 171 difference between gap sizes (ANOVA, $p = 0,171$, Fig. 3A). EP presented an average PAI (0.79

172 cm year⁻¹) four times greater than SRIL (0.19 cm year⁻¹, ANOVA, $p < 0.001$, Fig. 3B) and
 173 TNER had PAI (0.5 cm year⁻¹) almost threefold larger than SRIL (ANOVA, $p < 0.001$, Fig.
 174 3B). SRIL, TNER and EP presented statistical difference between them (ANOVA, $p < 0.001$,
 175 Fig. 3B). CEC 3 was more than twice (0.88 cm year⁻¹) higher than CEC 1 (0.42 cm year⁻¹,
 176 ANOVA, $p < 0.001$, Fig. 3C) and CEC 2 was almost twice than CEC 1 (0.42 cm year⁻¹,
 177 ANOVA, $p < 0.001$, Fig. 3C). The crown exposure class 1, 2 and 3 presented statistical
 178 difference between them (ANOVA, $p < 0.001$).

179 There was interaction of treatments versus crown exposure class (Two-way ANOVA, p
 180 < 0.001). CEC 1 presented the lowest means of the three treatments (0.14, 0.46, 0.64 cm year⁻¹,
 181 respectively to SRIL, TNER and EP). CEC 3 presented the highest means (0.43, 0.69, 0.95
 182 cm year⁻¹, respectively to SRIL, TNER and EP, Fig. 3D). EP showed the highest means in all
 183 crown exposure classes, and SRIL the lowest.



184

185

186

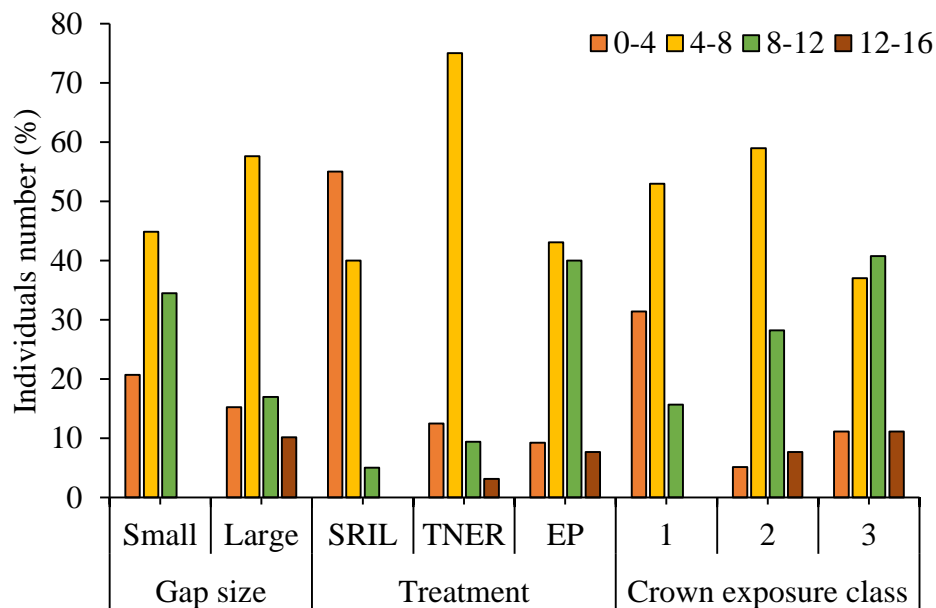
187

Figure 3: Periodic annual diameter (PAI) measured from diameter at breast height (DBH) at 1.3 m from the soil of each individual distributed in gap sizes, treatments (SRIL: Standard Procedures of RIL, TNER: Tending of

188 Natural Established Regeneration, EP: Enrichment Planting) and crown exposure classes (1: no direct light, 2:
 189 some lateral light, and 3: 10–90% overhead light), over 11 years in logging gaps in the managed forests of Jari
 190 S.A., Eastern Amazon, Brazil. In figure “D” lower cases indicate differences in treatments and upper case letters
 191 are for crown exposure class comparisons.

192

193 Large gaps, treatments where tending was applied (TNER and EP) and individuals with
 194 some lateral light and 10–90% overhead light (CEC = 2 and 3), reached the fourth diameter
 195 class (12-16 cm). SRIL, small gaps and the first crown exposure class did not pass of the third
 196 diameter class (8-12 cm).



197

198 Figure 4. Percentage of individuals per diameter class (0-4 cm; 4-8 cm, 8-12 cm and 12-16 cm) distributed in gap
 199 sizes, treatments and crown exposure classes, over 11 years in logging gaps in the managed forests of Jari S.A.,
 200 Eastern Amazon, Brazil.

201

202 4. Discussion

203 Growth was higher in enrichment planting (EP) and tending of natural regeneration
 204 (TNER) in logging gaps when compared to SRIL, which is the treatment that reflects the current
 205 procedures of RIL legally adopted in the Brazilian Amazon. There is no difference in growth
 206 of individuals according in relation to gap size, but according to their crown exposure class.
 207 Despite the high mortality of planted individuals (EP), in this treatment individuals had better
 208 performance in growth, especially those with higher crown exposure class (CEC). These
 209 outcomes are positive in order to increase the low natural densities of *D. excelsa* and to improve
 210 its performances in survival, growth and structure.

211 The better survival of *D. excelsa* in SRIL and TNER in comparison to EP is probably
 212 due to the acclimatation phase and even may be a result of the low quality of many planted

213 seedlings. The availability of high-quality seedlings of native commercial species has been a
214 serious bottleneck to permit enrichment planting in managed forests of the Brazilian Amazon
215 (Neves et al., 2019). For the success of enrichment planting in logging gaps, it is recommended
216 a rigorous seedling production, with fertilizers, as well as annual cleanings once seedlings
217 planted in the field normally face strong competition (Gomes et al., 2019).

218 Was observed after RIL over time the occurrence of new logging gaps. According
219 Dionisio et al. (2017), mortality rates increase in the first five years after logging and the effects
220 of RIL remain up to seven years after RIL. Therefore, these logging gaps that are created over
221 time could be used to apply post-harvesting silvicultural treatments, after all, the entire structure
222 of the RIL are recent created, such as the drag roads, and are available for tending and
223 enrichment planting, optimizing the species conservation and the timber production. To achieve
224 this, it is necessary the development of new public policies to improve the current forest
225 management regulations in the Brazilian Amazon (Neves et al., 2019).

226 It was remarkable the occurrence of high natural regeneration in the annual production
227 unit (APU). Also noteworthy, based on observations of logging gaps and their surroundings,
228 significant amount of *D. excelsa* regeneration in the following years after the RIL. We suppose
229 that there was a disturbance in the past in which an adequate edaphoclimatic condition was
230 created for this high occurrence. The high mortality observed in the first four years strongly
231 decreased in EP. So the density will be higher if combined the techniques of: (a) replanting
232 seedlings from the nursery; (b) after RIL, logging gaps as well as skid trails, and log decks can
233 be used, aiming at saving nursery costs and mitigating the effects of the seedling rustification
234 process; and (c) tending the seedlings naturally present in gaps, these so-called complementary
235 trees (Neves et al., 2019).

236 *Dinizia excelsa* presented good growth performance where tending was applied
237 however, there was no difference in increment by gap sizes. This finding contradicted what was
238 expected, a high diameter growth of trees in larger gaps. The more sunlight incidence inside
239 logging gaps, the better is trees growth, which fact explains the higher basal area of treatments
240 with tending (TNER and EP). Even small logging gaps could be used to grow *D. excelsa* by
241 forestry companies, which can represent economic advantages in terms of silvicultural costs.
242 Regarding the silvicultural treatment, TNER presented the highest cost-benefit relation
243 (Schwartz et al., 2016), so it can be widely applied under lower cost in managed forests rich in
244 natural regeneration of commercial species, as can be observed in the present study.

245 Sunlight is the main limiting factor for plant growth and survival, so in evolutionary
246 terms, it determines the plant's survival strategy in a forest (Begon et al., 2009; Odum, 2006).

247 Light-demanding species, such as *D. excelsa*, tolerate certain shading levels, however, they
248 grow faster under overhead light, as showed in this current study. The increment of the EP with
249 CEC 3 performed better due to the high mortality of the other planted species (Neves et al.,
250 2019; Schwartz et al., 2013), generating less competition, allied to the fact that there was the
251 silvicultural treatment of tending (competition cleaning) in these gaps. As in TNER there is low
252 mortality (Neves et al., 2019; Schwartz et al., 2013), there is a greater number of individuals
253 competing in the area, so this justifies its lower growth compared to EP that had the highest
254 mortality and highest growth. Successful enrichment planting in logging gaps confirms the
255 results found in other experiments worldwide with enrichment planting in gaps (Doucet et al.,
256 2009; Gomes et al., 2019, 2010; Lopes et al., 2008; Neves et al., 2019; Quédraogo et al., 2014;
257 Schwartz et al., 2013; Taffarel et al., 2014; Vieira et al., 2018). These studies reinforce the
258 efficiency of enrichment planting in logging gaps, which comes as a viable silvicultural
259 alternative for managing tropical forests.

260 Individuals with some lateral light and 10–90% overhead light (CEC 2 and 3) of the
261 treatments where tending was applied (TNER and EP) reached the fourth diameter class (12-16
262 cm) and presented the highest means of PAI (cm year^{-1}). SRIL and CEC 1 were not higher than
263 the third diameter class (8-12 cm) and presented the lowest means of PAI. This positive effect
264 reinforces the importance of improving post-harvesting silvicultural treatments in forest
265 management plans. Besides this, the positive outcomes could be extrapolated and tested in
266 species of the same ecological group. For example, species of the same ecological group under
267 TNER and EP treatments and with better sunlight incidence in canopy logging gaps can shorten
268 the time required to recover timber losses caused by RIL (Neves et al., 2019) and help to
269 mitigate the delay in several tropical forests to recover harvested stocks (Hu et al., 2020; Shima
270 et al., 2018).

271 The dominance of *D. excelsa* in treatments with tending was twice higher than the
272 standard procedures of RIL. Schwartz et al. (2016) found the best cost-benefit relation for
273 TNER, but the authors considered a set of species in their study, which included some slow-
274 growth species. In this study we focused on the light-demanding species *D. excelsa*, which
275 presented a high silvicultural performance.

276

277 **5. Conclusion**

278 Individuals of *Dinizia excelsa* presented better survival and growth under post-
279 harvesting silvicultural treatments of tending both planted and naturally regenerated

280 individuals. Growth of individuals did not respond to differences in gap sizes, but according to
 281 crown exposure, individuals with overhead light has higher growth.

282

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288

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1 **Silviculture of a very common commercial timber species in the Amazon: managing**
 2 ***Tachigali glauca* in logging gaps²**

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10

11 **Abstract**

12 The objective of this paper was analyze the medium-term effects of post-harvesting
 13 silvicultural treatments over growth and survival of *Tachigali glauca* individuals planted or
 14 naturally established in canopy logging gaps. under tending and enrichment planting with
 15 different crown exposure intensities. A total of 181 *Tachigali glauca* individuals were used in
 16 the experiment. Part of them was naturally present while other was planted in 26 gaps created
 17 by tree felling in order to evaluate the specie medium-term responses to the following
 18 treatments: (1) standard procedures of RIL (SRIL) or control, (2) tending of the naturally
 19 established regeneration (TNER) and (3) enrichment planting (EP). Effects of the EP
 20 silvicultural treatment were positive to *T. glauca*, since it presented high survival and growth
 21 of the trees present in canopy logging gaps. Furthermore, EP presented PAI value almost four
 22 times greater than SRIL, which is the treatment that reflects the current procedures required for
 23 the employment of reduced impact logging (RIL) in the Brazilian Amazon. The EP growth is
 24 reflected in percentage of individuals in the fourth diameter class (15-20) and means that EP
 25 trees are growing faster than trees of the other two treatments.

26 **Keywords:** assisted densification; conservation and timber production; sustainable forest
 27 management; tropical forest.

28

29 **1. Introduction**

30 “Can timber provision from Amazonian production forests be sustainable?”. Pioniot et
 31 al. (2019) approached the question with simulations finding that, regardless the cutting cycle
 32 duration and logging intensities, selectively logged forests are unlikely to meet timber demands
 33 over the long term as timber stocks are predicted to steadily decline. Moreover, 40–50 years is

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34 not time enough for mixed dipterocarp tropical forests in Kalimantan, Indonesia to recover
35 original species composition after logging (Shima et al., 2018).

36 Some alternatives to mitigate such tropical forest issues are the application of post-
37 harvesting silvicultural treatments in canopy logging gaps, as tending and enrichment planting.
38 These treatments bring positive outcomes to reach more sustainable future cutting cycles
39 (Doucet et al., 2009; Gomes et al., 2019, 2010; Lopes et al., 2008; Neves et al., 2019; Quédraogo
40 et al., 2014; Schwartz et al., 2013; Taffarel et al., 2014; Vieira et al., 2018). Tending consists
41 in the liberation of target individuals against competing tree species and lianas, providing better
42 light availability to improve survival and more rapid growth (Brokaw 1985, Brown & Whitmore
43 1992).

44 Higher light availability promoted by tending tends to benefit mainly individuals of
45 pioneer and light-demanding species, as the case of *Tachigali glauca* Tul. in three managed
46 forests sampled by Schwartz et al. (2017a). *T. glauca* is an monocarpic (Foster, 1977), fast-
47 grow and light-demanding endemic tree species in Brazil that occurs naturally not only in the
48 North region of Brazil (states of Acre, Amazonas, Amapá, Pará, Rondônia and Roraima), but
49 also in its Northeast (state of Maranhão) and Midwest regions (state of Mato Grosso) (Lima,
50 2015). To Pará state, in 10 years (01/Jan/2006-21/Feb/2016), 14,415.92 m³ of *T. glauca* timber
51 was traded by 453.012,10 USD (1,954,656.46 BRL, rate of 11/Mar/2020), with an average price
52 of 31.29 USD (“SEMAS - Sisflora PA,” n.d.). Schwartz et al. (2016), in a profitability analysis
53 of a set of species, including *T. glauca*, found the best cost-benefit relation in the treatment of
54 tending of the naturally established regeneration. So, the treatments will improve the wood trade

55 In the current Sustainable Forest Management Plans (SFMP) in the Brazilian Amazon,
56 almost no investments have been addressed to post-harvest treatments. In this sense, new public
57 policies to improve the current forest management regulations in the Brazilian Amazon (Neves
58 et al., 2019) using commercial species as *T. glauca* can call the forest managers attention to

59 invest in silvicultural techniques, especially over pioneer and light-demanding species. Thus,
60 the objective of this paper was to analyze medium-term effects of post-harvesting silvicultural
61 treatments over growth and survival of *T. glauca* individuals planted or naturally established in
62 canopy logging gaps under tending and enrichment planting with different crown exposure to
63 sunlight. According to the *T. glauca* characteristics of be a fast-growing and light-demanding
64 tree species, it is hypothesized that, after application of silvicultural treatments on logging gaps,
65 the individuals will respond positively in growth and survival.

66

67 **2. Materials and methods**

68 *2.1. Study area*

69 Data were collected from a field experiment carried out in the forest management area of
70 the forestry company Jari Florestal SA under the project ‘Logging Gaps Management’
71 coordinated by Embrapa Eastern Amazon in cooperation with Jari Florestal S.A. The study area
72 is located in the Jari valley, Almeirim municipality (1° 9' S, 52° 38' W), Pará state, Brazil.
73 Average annual precipitation is 2200 mm and the annual average temperature is 26 °C. The
74 most common vegetation type is ombrophilous dense forest popularly as *terra firme* forest,
75 where the most dominant soils are yellow latossols (Azevedo, 2006).

76 Jari Florestal has a total area under forest management of 545,535 ha and all harvesting
77 operations follow RIL techniques. A total of 181 seedlings and saplings of *Tachigali glauca*
78 were used in the experiment. Part of them was naturally present while others were planted in
79 26 gaps created by tree felling in order to evaluate the specie medium-term responses to the
80 following treatments: (1) standard procedures of RIL (SRIL) or control, (2) tending of the
81 naturally established regeneration (TNER) and (3) enrichment planting with tending the planted
82 seedlings (EP). All seedlings and saplings had 40-80 cm in height.

83

84 *2.2. Experimental design*

85 The experiment was established in 2006 and 2007 in the logging compartments harvested
86 in 2004 and 2006. In SRIL, marked individuals were only monitored, with no additional
87 silvicultural treatments, according to the current forest management regulations in the Brazilian
88 Amazon. In the other two treatments, post-harvesting silvicultural procedures were applied in
89 addition to all steps required to employ RIL.

90 Tending consisted in the liberation of target individuals against competing individuals of
 91 tree species, commonly pioneers, and lianas. This treatment was applied over seedlings and
 92 saplings of commercial tree species naturally established (TNER) and planted (EP) in all
 93 measurement years. EP treatment was established in 1-year-old gaps with complete logging
 94 residual removal for further energy production by the forestry company and every seedling was
 95 planted in spacing of 2.5 m × 2.5 m inside the logging gaps.

96 According to the pre-logging inventory (PL) done in 2005, as the standard procedures of
 97 RIL, the annual production unit (APU) had an area of 7,600 ha and 6,555 individuals of *T.*
 98 *glauca* individuals, all above 35 cm in DBH. Based on this information, it was possible to
 99 calculate the pre-logging absolute density (ADe), dominance (ADo) and basal area (G) of *T.*
 100 *glauca*. Forest characteristics as volume, species composition, diameter structure, and diversity
 101 indexes of the study area (annual production unit) are available in Lopes De Souza et al. (2014),
 102 Neves et al. (2019), and Schwartz et al. (2013). The dendrological variables for treatments were
 103 based on data collected in 2018.

104 In 2018, all trees that reached ≥ 300 cm in height in 2012 were included in the analyses,
 105 totalizing 31, 16 and 115 trees in SRIL, TNER and EP, respectively (Table 1). This means that
 106 every planted or naturally regenerated seedling < 300 cm in height in 2012 was not considered
 107 in this study.

108

109 Table 1: Number of individuals (N), basal area (G), absolute density (ADe) and dominance (ADo) of *Tachigali*
 110 *glauca* in the forest management area of Jari SA, Eastern Amazon, Brazil. PL (pre-logging), SRIL (standard
 111 reduced impact logging/control), TNER (tending of the naturally established regeneration), and EP (enrichment
 112 planting).

	PL*	SRIL	TNER	EP
Initial measurement year	2005	2006	2006	2006
Number of individuals in the initial measurement year	6555	35	17	129
Number of individuals in the last data collect	uninformed	31	16	115
Area (ha)	7600	0.593	0.659	0.604
G (m ² ha ⁻¹)	0.15	0.10	0.14	1.77
ADe (ind ha ⁻¹)	0.9	57.4	25.8	190.4
ADo (m ² ha ⁻¹)	0.64	0.1	0.21	0.43

113 * The *T. glauca* Pre-logging (PL) information in 2005.

114

115 Mortality rate (m) was calculated through the formula “ $m = 1 - (Nt2 / Nt1)(1/t)$ ”, where
 116 Nt1 = Number of live trees in the first sampling, Nt2 = number of trees that survived until the
 117 second sampling and t = years between first and second sampling (Sheil et al., 1995). The

118 logging gap area was calculated by the ellipse formula. The basal area was the sum of tree cross
119 sections of each treatment. Each tree cross section was obtained with the formula $g = \pi * (DBH/2)^2$
120 in square meters per hectare, where DBH = diameter at breast height at 1.3 m from
121 the soil.

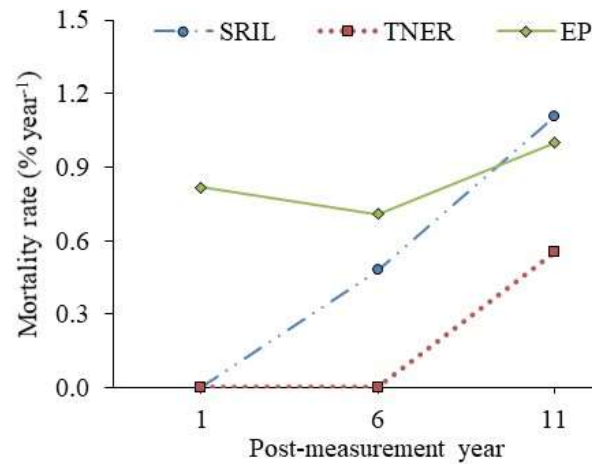
122 Periodic annual diameter (PAI) measured from the DBH of each individual was
123 calculated using the formula $PAI = ((DBHt2 - DBHt1)/n)*365$, where DBHt1 = individual's
124 diameter at the initial sampling, DBHt2 = individual's diameter at the final sampling, and n =
125 days between first and second sampling. The 117 individual PAI of *T. glauca*, which reached
126 minimum height of 300 cm by 2012, were compared by treatment and crown exposure class
127 (CEC). The individuals crown exposure class were assigned with the system of Clark & Clark
128 (1992): CEC 1 (no direct light, 8 individuals), CEC 2 (some lateral light, 37 individuals) and
129 CEC 3 (10–90% overhead light, 101 individuals).

130 Once applied the Shapiro-Wilk test on PAI by treatment and crown exposure class, the
131 data did not present a residual normal distribution ($p \text{ normal} < 0.05$), which required a square
132 root transformation, before running ANOVA and the post-hoc Tukey's pairwise test to
133 treatments and crown exposure class. The number of individuals, in percentage, were set in four
134 diameter class ranging 5 cm in DBH from 0 up to 20 cm.

135 The PAI of each individual present in the eight gaps of the EP treatment was also
136 compared by ANOVA with post-hoc Tukey's pairwise test. The area, in square meters, of the
137 eight logging gaps was: EP1 = 827.81 m²; EP2 = 785.4 m²; EP3 = 1278.63 m² (the largest gap);
138 EP4 = 632.25 m²; EP5 = 572.56 m²; EP6 = 523.08 m² (the smallest gap); EP7 = 643.24 m² and
139 EP8 = 775.97 m²). All analyses were performed using the R version 3.0.2 (2016).

140 **3. Results**

141 Mortality rates of all treatments were lower than 2%, with only one death in TNER
142 treatment (0.55% year⁻¹). The highest mortality rate was to SRIL treatment (1.10% year⁻¹,
143 Figure 1).

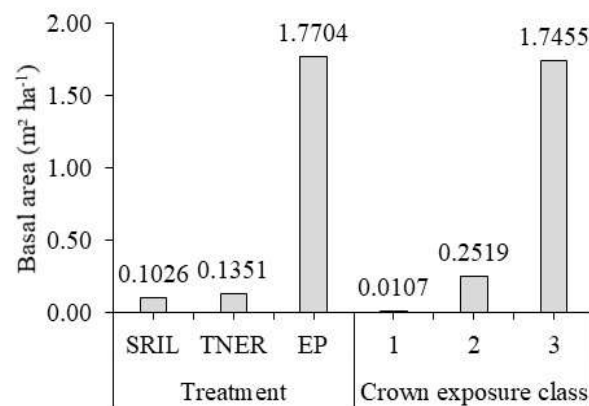


144

145 **Figure 1.** Mortality rates of standard procedures of reduced-impact logging (SRIL), tending of the naturally
 146 established regeneration (TNER) and enrichment planting (EP) treatments by one, six and 11 years in logging gaps
 147 in the managed forests of Jari company, Eastern Amazon, Brazil.

148

149 The highest basal area was observed in EP treatment (1.7704 m² ha⁻¹) and in individuals
 150 under crown exposure class 3 (1.7455 m² ha⁻¹). The lowest basal area was observed in SRIL
 151 treatment (0.1026 m² ha⁻¹) and in CEC 1 (0.0107 m² ha⁻¹, Figure 2).



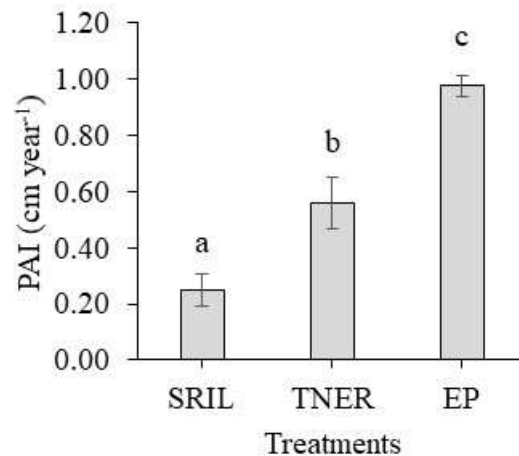
152

153 Figure 2: Basal area (m² ha⁻¹) of treatments (SRIL: Standard Procedures of RIL, TNER: Tending of Natural
 154 Established Regeneration, EP: Enrichment Planting) and crown exposure classes (1: no direct light, 2: some lateral
 155 light, and 3: 10–90% overhead light), over 11 years in logging gaps in the managed forests of Jari, Eastern Amazon,
 156 Brazil.

157

158 Considering PAI of each treatment, the EP mean (0.98 cm year⁻¹) was almost four times
 159 greater than SRIL (0.25 cm year⁻¹, ANOVA, $p < 0.001$); and TNER (0.56 cm year⁻¹) was twice
 160 greater than SRIL (0.25 cm year⁻¹, ANOVA, $p < 0.001$). SRIL, TNER and EP presented
 161 statistical differences among them (ANOVA, $F_{2;143} = 56.76$, $p < 0.001$, Figure 3).

162



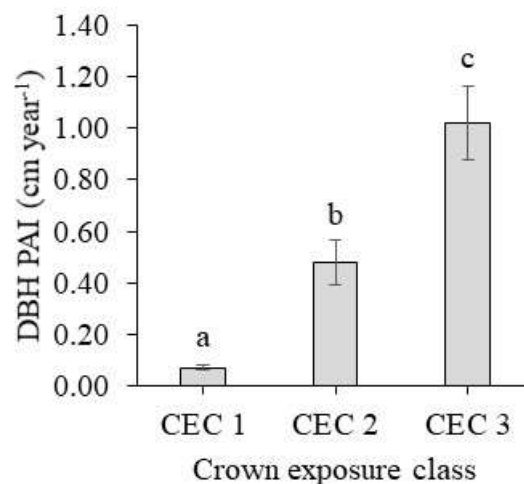
163

164 Figure 3: Periodic annual diameter (PAI) mean (\pm SE) of the diameter at breast height (DBH) at 1.3 m from the
 165 soil of each individual per treatment over 11 years in logging gaps in the managed forests of Jari, Eastern Amazon,
 166 Brazil. Letters indicate differences in ANOVA with post-hoc Tukey's pairwise test.

167

168 PAI (cm year⁻¹) in relation to the crown exposure class (CEC), the treatments presented
 169 a substantial difference among their means (ANOVA, $F_{2,143} = 83.34$, $p < 0.001$, Figure 4). CEC
 170 3 (1.02 cm year⁻¹) was almost 15 times higher than CEC 1 (0.07 cm year⁻¹, ANOVA, $p < 0.001$)
 171 and CEC 2 (0.48 cm year⁻¹) was six times higher than CEC 1 (0.07 cm year⁻¹, ANOVA, $p <$
 172 0.001; Figure 4).

173



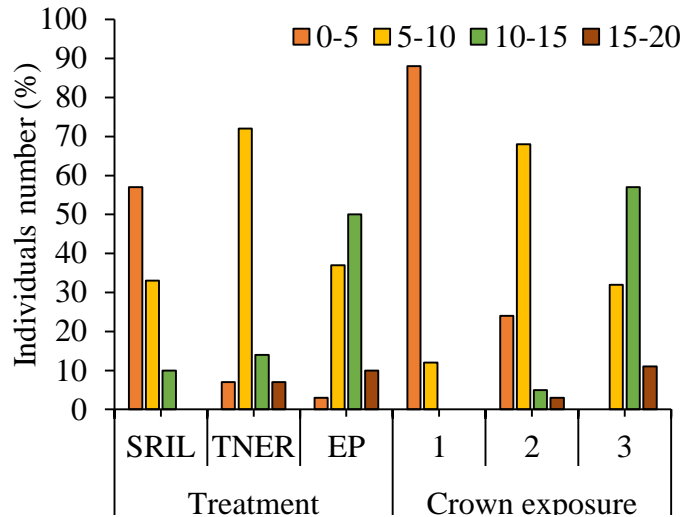
174

175 Figure 4: Periodic annual diameter (PAI) mean (\pm SE) of the DBH at 1.3 m from the soil of each individual
 176 distributed in crown exposure classes over 11 years in logging gaps in the managed forests of Jari, Eastern Amazon,
 177 Brazil. Letters indicate differences in ANOVA with post-hoc Tukey's pairwise test.

178

179 Individuals under tending treatments (TNER and EP) and with some lateral light and

180 10–90% overhead light (2 and 3 crown exposure class), reached the fourth diameter class (15–
 181 20 cm), while SRIL individuals and those with CEC 1 did not pass over the third diameter class
 182 (10–15 cm, Figure 5).

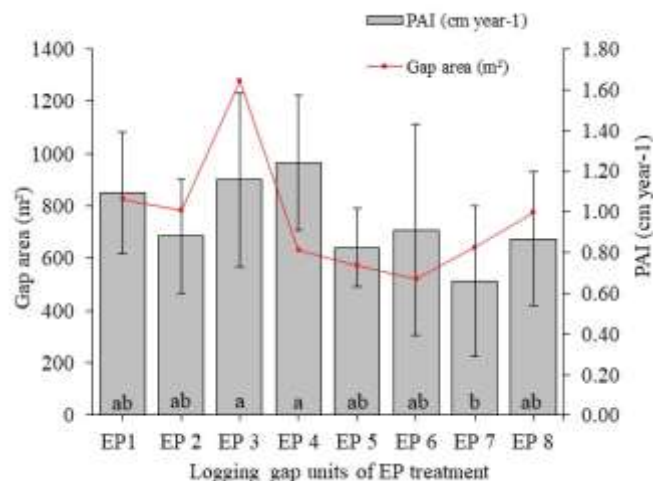


183

184 **Figure 5.** Percentage of individuals per diameter class (0-5 cm; 5-10 cm, 10-15 cm and 15-20 cm) distributed in
 185 treatments (SRIL: Standard Procedures of RIL, TNER: Tending of Natural Established Regeneration, EP:
 186 Enrichment Planting) and crown exposure classes (1: no direct light, 2: some lateral light, and 3: 10–90% overhead
 187 light), over 11 years in logging gaps in the managed forests of Jari, Eastern Amazon, Brazil.

188

189 The logging gap EP3 (area = 1,278.63 m²) and EP4 (area = 632.25 m²) presented the
 190 highest PAI means (1.16 and 1.24 cm year⁻¹ respectively). The logging gap EP7 (area = 643.24
 191 m²) presented the lowest PAI mean (0.66 cm year⁻¹). There were statistical differences between
 192 EP 7 from EP4 and EP3 (ANOVA, $F_{7,103} = 4.16$, $p < 0.01$, Figure 6).



193

194 **Figure 6.** Mean (\pm SD) of periodic annual increment (PAI cm year⁻¹) and Gap area (m²) of the eight logging gap
 195 units of enrichment planting treatment in the managed forests of Jari, Eastern Amazon, Brazil. Letters indicate
 196 differences in ANOVA with post-hoc Tukey's pairwise test.

197

198

199 4. Discussion

200 Effects of the silvicultural treatment enrichment planting (EP) were positive to *Tachigali*
201 *glauca*, since it presented high survival and growth of the trees present in canopy logging gaps.
202 EP presented PAI value almost four times greater than SRIL, which is the treatment that reflects
203 the current procedures required for the employment of reduced impact logging (RIL) in the
204 Brazilian Amazon. The EP growth also showed as the percentage of individuals in the fourth
205 diameter class (15-20), where EP trees grew much faster than trees of the two other treatments.

206 For the success of enrichment planting in logging gaps it is essential the production of
207 high quality seedlings, with fertilizers, as well as annual cleanings once seedlings planted in the
208 field normally face strong competition (Gomes et al., 2019; Schwartz et al., 2013). However,
209 in the present study the seedlings possibly had low quality and was not utilized fertilizer, only
210 annual cleanings were applied, therefore, it was expected low survival in enrichment planting
211 treatments of canopy logging gaps (Neves et al., 2019) mainly due to the rustification phase,
212 (Adenesky-Filho et al., 2017; Neves et al., 2019), which is the phase in which the plant takes
213 to adapt to the new edaphoclimatic conditions. However, the EP treatment presented low
214 mortality rate, which may be a positive effect of the silvicultural treatment.

215 Mortality rates vary among species and diameter classes, however, the average mortality
216 rate for the three treatments (0.88% year⁻¹), with an emphasis on TNER treatment with only
217 one dead tree (0.55 year⁻¹), is within the range of values reported in tropical forests in the
218 Amazon and Southeast Asia, with annual mortality rates ranging between 1% and 2% per year
219 in unexplored natural tropical forests (Swaine et al., 1987), with values generally higher for
220 recently harvested forests and decreasing over time (Dionisio et al., 2017; Dionisio et al., 2018).
221 It is expected mortality rate of individuals lower than 20% in forests under silvicultural
222 treatments. For example, Neves et al. (2019) reported mortality rates of individuals of
223 commercial species ranging from 3% to 10% per year in forests under different silvicultural
224 treatments at the Jari company 11 years after harvest. Das Chagas et al. (2012) and Gomes et
225 al. (2010) reported mortality rates of 19% and 14% of planted individuals in logging gaps five
226 and one year after harvesting, respectively. The highest mortality rates are observed during the
227 first period of two to seven years after RIL (Shenkin et al., 2015; Dionisio et al., 2018), and the
228 mortality rates stabilize afterwards, between 5 years and 10 years (Dionisio et al., 2018, 2017;
229 Sist et al., 2003) or after 15 years post-harvest (Avila et al., 2017).

230 It is expected over time after RIL the opening of new canopy gaps by the tree fall of
231 remaining trees. According Dionisio et al. (2017) and Sist et al. (2003), the mortality rate of
232 remaining trees is higher in the first five years after logging and its effects remain up to 10

233 years. These new gaps could also be used in the application of post-harvesting silvicultural
234 treatments, after all, there is the entire structure of the RIL, such as the drag roads, optimizing
235 the species conservation and the timber production.

236 EP presented a basal area 13 times greater or 92% larger than the standard procedures
237 of RIL, which is a positive effect of the treatment of *T. glauca* in logging gaps. These results
238 were only possible due to the increase in planting density. Neves et al. (2019) also observed
239 that the enrichment with commercial species on logging gaps by the company Jari, significantly
240 increased the basal area of individuals. The authors reinforce that these treatments bring positive
241 results to achieve future more sustainable cutting cycles in the Brazilian Amazon.

242 Light intensity had also positive influence over the increase growth in diameter. The
243 evolutionary strategy of light-demanding species, such as *T. glauca*, consists of a greater
244 investment in secondary growth (diameter) than primary growth (height), due to the
245 photosynthetically active radiation that reaches the forest floor. Such process becomes more
246 intense as it increases the canopy opening through disturbances caused by human activities such
247 as selective logging (Jardim et al., 2007; Reis et al., 2014) and silvicultural treatments (Gomes
248 et al., 2010; Neves et al., 2019; Schwartz et al., 2017b, 2013; Vieira et al., 2018).

249 Radiation is the main limiting factor for plant survival, so it determines the plant's
250 survival strategy in a forest (Begon et al., 2009; Odum, 2006). Light-demanding species, such
251 as *T. glauca*, present better development in environments with greater sunlight incidence, as
252 was found in the current study. The PAI of EP and crown exposure class number three (CEC
253 3) performed better due to the high mortality of the other planted species (Neves et al., 2019;
254 Schwartz et al., 2013), diminishing competition, allied to the fact that there was the tending in
255 these gaps. As in TNER there is low mortality (Neves et al., 2019; Schwartz et al., 2013), there
256 is a greater number of natural regeneration competing in the area, so this justifies its lower
257 growth in diameter compared to the EP, but is needed an evaluation of others factors as growth
258 in height for example.

259 Forest disturbances which result in canopy logging gaps, can be used as an effective
260 way to conserve rare tree species (low density of individuals) and / or those with low natural
261 regeneration and slow growth (Neves et al., 2019). According to Schwartz and Lopes (2015)
262 increasing the density of individuals of rare species in logging gaps can function as an artificial
263 refuge for endangered species, maintaining their genetic diversity. This is possible through
264 assisted densification (artificial increase in the number of individuals per unit area of tree
265 species in their own natural habitats), a procedure that can help to ensure the third cutting cycle
266 (Dionisio et al. (2017).

267 Successful enrichment planting in logging gaps confirms the results found in other
268 experiments worldwide with enrichment planting in gaps (Doucet et al., 2009; Gomes et al.,
269 2019, 2010; Lopes et al., 2008; Neves et al., 2019; Quédraogo et al., 2014; Schwartz et al.,
270 2013; Taffarel et al., 2014; Vieira et al., 2018). These studies corroborate the efficiency of
271 enrichment planting in logging gaps, which comes as a viable silvicultural alternative for
272 managing tropical forests. Regarding the silvicultural treatment, TNER presented the highest
273 cost-benefit relation (Schwartz et al., 2016), so it can be widely applied under low cost in
274 managed forests rich in natural regeneration of commercial species, as the case of the study
275 area in the Jari's managed forests.

276 Individuals with some lateral light and 10–90% overhead light (2 and 3 crown exposure
277 class) of the treatments where tending was applied (TNER and EP) reached the fourth diameter
278 class (15-20 cm) and presented the highest PAI (cm year^{-1}). SRIL and the first crown exposure
279 class (CEC 1) had no individuals above third diameter class (10-15 cm) and presented the
280 lowest means of PAI (cm year^{-1}), this was expected because *T. glauca* is a light-demanding
281 specie. This positive effect reinforces the importance of improving post-harvesting silvicultural
282 treatments in forest management plans. Besides this, its positive outcomes can be expanded and
283 tested in species of the same ecological group. For example, species of the same ecological
284 group subjected to TNER and EP treatments allied to better sunlight incidence in canopy
285 logging gaps could shorten the time required to recover losses caused by RIL (Neves et al.,
286 2019) and help to mitigate the delay in several tropical forests to recover harvested stock (Hu
287 et al., 2020; Shima et al., 2018).

288 The dominance of *T. glauca* in EP treatment ($0.43 \text{ m}^2 \text{ ha}^{-1}$) is four times greater than
289 dominance of SRIL, TNER ($0.21 \text{ m}^2 \text{ ha}^{-1}$) is twice than SRIL ($0.10 \text{ m}^2 \text{ ha}^{-1}$). But in EP, *T.*
290 *glauca* was planted with only four other species having a high density in the gap, while in the
291 other treatments, its density was much lower. Schwartz et al. (2016) found the best cost-benefit
292 relation for TNER, but the authors considered a set of species in their study, which included
293 some slow-growth species. In this study we focused on the light-demanding species *T. glauca*,
294 which presented a high silvicultural performance.

295

296 **5. Conclusions**

297 *Tachigali glauca* presented better survival and growth in the highest crown exposure
298 classes and in canopy logging gaps under enrichment planting treatment in comparisons to
299 lowest crown exposure classes and in gaps under standard procedures of reduced impact

300 logging treatment.

301

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307

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CONCLUSÕES E SUGESTÕES PARA PRÓXIMOS ESTUDOS

De um modo geral nós podemos perceber que os tratamentos silviculturais pós-colheita influenciam positivamente na sobrevivência e crescimento de espécies comerciais da Amazônia. Algumas implicações/potencialidades dos tratamentos pós-colheita em clareiras são:

- Aliar produção madeireira com a conservação de espécies de baixa densidade da Amazônia, principalmente as demandantes de luz;
- Os nossos resultados a médio prazo, demonstraram que podemos produzir e/ou conservar em um espaço de tempo menor que o que acontece naturalmente após o preconizado hoje na Exploração de Impacto Reduzido (EIR);
- Há a possibilidade de utilizarmos as clareiras como pomar de sementes e mudas;
- Há a possibilidade de uso dessas clareiras como zonas de produtos florestais não madeireiros;
- Conforme outro trabalho do mesmo projeto, foi mostrado que a EIR gera mortalidade em até 11 anos após a exploração, por isso há ainda clareiras que são resquícios de ações antrópicas que podem ser utilizadas, afinal já há toda a estrutura de ramais, pátios, clareiras etc, só precisa ser incorporado os tratamentos para que haja essa otimização do uso de clareiras provenientes da EIR;
- Apresenta potencial econômico, principalmente com espécies de crescimento rápido como o Paricá;
- Podemos aproveitar para criação de zonas como Banco Ativo de Germoplasma *in situ*, onde iremos dedicar às espécies de baixa densidade na área e/ou aquelas ameaçadas ou em risco;
- Na área de estudo temos ensaios que envolvem tanto a floresta primária quanto área degradada pela extração de cascalho, por exemplo, então estas técnicas podem ser estudadas para aplicação também em florestas secundárias e/ou áreas degradadas;
- A metodologia pode ser aplicada preferencialmente no período pré-chuvoso ou chuvoso, este fato irá favorecer aos trabalhadores que, geralmente, são demitidos no período de embargo.